Blood Methylmercury and Fish Consumption Among People of Childbearing Age in the General U.S. Population

NHANES 1999-March 2020

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Executive Summary

In the United States, exposure to methylmercury (MeHg) in humans occurs largely through the consumption of fish (National Research Council, 2000; Rice *et al.*, 2000). Blood concentrations of MeHg in women of childbearing age are of particular interest because exposure to MeHg *in utero* is associated with adverse health effects, e.g., neurodevelopmental deficits such as IQ and motor function deficits in children (Mergler *et al.*, 2007; National Research Council, 2000).

This report documents an analysis of fish consumption and blood mercury concentrations in women aged 16-49 years in the United States using data collected by the National Health and Nutrition Examination Survey (NHANES) from 1999 through March 2020 (10 survey releases). NHANES is a continuous survey designed to collect data on the health and nutritional status of the U.S. population. The NHANES reports data on chemicals, or their metabolites as measured in blood and urine samples collected from a statistically representative sample of the U.S. population. CDC releases NHANES data every two years.

In this study, we applied fish tissue mercury concentrations to fish species reported being consumed by the study participants, and estimated the usual intake of mercury through fish consumption (United States Environmental Protection Agency, 2014). We imputed blood MeHg from blood total and blood inorganic mercury data and investigated the trend of blood MeHg concentration over time by demographic characteristics. Additionally, we looked the association between blood mercury concentration and demographic characteristics and the association between blood mercury concentration and estimated usual intake of mercury. We also looked for trends in frequency of fish consumption as well as the association of fish consumption and mercury intake with demographic characteristics. Finally, we looked for geographic differences in blood mercury.

Key Findings

Trends in MeHg concentrations in blood: There are statistically significant decreasing trends in blood MeHg concentrations over time (higher in 1999-2000 and lowest in 2017-March 2020) and by demographic characteristics (such as education, income, race/ethnicity, and age over the study period).

- The geometric mean blood MeHg concentration in 1999-2000 survey release is 1.94 times higher than the geometric mean in 2017-March 2020 survey release, representing a 48 percent decrease between NHANES 1999-2000 (the earliest set of data analyzed) and 2017-March 2020 (the most recent dataset).
- The percentage of women of childbearing age with blood MeHg concentrations over 5.8 μ g/L in 1999-2000 is about 3.5 times the concentration found in 2001-March 2020, representing a 71 percent decrease. There is a statistically significant difference between the survey releases for the percent with blood MeHg concentrations over 5.8 μ g/L (Rao-Scott Chi-square p<0.001). No significant difference (p=0.12) is found between the survey releases after removing survey release 1999-2000.
- The geometric mean blood MeHg concentration in women who reported their race as "Other Race Including Multi-Racial" (which includes Asian, Native American, Pacific and Caribbean Islander, Alaska Native, multiracial, and unknown race) in 1999-2004 is 1.82 times higher than the geometric mean in 2017-March 2020 data, representing a 45 percent decrease between NHANES 1999-2004 and 2017-March 2020.
- Higher blood MeHg concentrations are observed with increasing age, ratio of family income to poverty, education level, and among participants who reported their race as "Other Race Including Multi-Racial."

Predictors of MeHg concentrations in blood: Transformed usual intake (TUI) is a significant predictor (p<0.0001) of blood MeHg concentrations. The rate of increase in blood MeHg concentration due to TUI varied by education, race/ethnicity, and log-transformed body weight. Other significant predictors of blood MeHg concentrations are NHANES survey release, education, race/ethnicity, log-transformed hematocrit concentration, and log-transformed bodyweight. Household income is a marginally significant predictor of blood MeHg concentrations (p=0.054).

Trends in fish consumption by demographic characteristics and geography: Blood MeHg

concentrations are positively associated with the reported frequency of fish consumption over the previous 30 days. While significant differences in reported frequency of fish consumption are found across the six NHANES survey releases from 1999-2010 in a previous study (United States Environmental Protection Agency, 2013), there are no statistically significant differences in reported frequency of fish consumption across the more recent NHANES survey releases in the current study (2013 to March 2020). There are statistically significant differences in blood MeHg concentrations geographically, with higher levels among residents of coastal counties compared to residents of non-coastal counties. The residents of the Northeast region have the highest levels of the four regions, followed by the West, South, and the Midwest.

Overarching Interpretation

There is a decreasing trend in the geometric mean of blood MeHg concentrations. While the estimated amounts of total fish eaten over the previous 30 days in NHANES 2013-March 2020 are at the higher end of estimated amounts for NHANES 1999-2010, the estimated mercury intake from total fish consumption over the previous 30 days in NHANES 2013-March 2020 are lower than estimated intakes for NHANES 1999-2010. This suggests that women of childbearing age are potentially choosing to eat fish that tend to have lower mercury concentrations leading to lower estimates of mercury intake per unit body weight.

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List of Abbreviations

- CDC Centers for Disease Control and Prevention
- DK Don't know
- DL Detection limit
- EPA Environmental Protection Agency
- FDA Food and Drug Administration
- FNDDS Food and Nutrient Database for Dietary Studies
- IHg Inorganic mercury
- MeHg Methylmercury
- NCHS National Center for Health Studies
- NCI National Cancer Institute
- NHANES National Health and Nutrition Examination Survey
- RR Relative Ratio
- THg Total mercury
- TUI Transformed usual intake
- USDA United States Department of Agriculture

BACKGROUND AND PURPOSE 1

In the United States, exposure to methylmercury (MeHg) in humans occurs largely through the consumption of fish (National Research Council, 2000; Rice et al., 2000). Mercury released into the environment is converted to MeHg in sediments and in the water column and bioaccumulates through aquatic food webs. This bioaccumulation leads to increased levels of MeHg in larger, older, predatory fish; concentrations in fish tissue may exceed a millionfold the concentrations in water (National Research Council, 2000). Fish and shellfish tissue contaminated by MeHg can put human health at risk. Blood total mercury (THg) concentrations reflect exposure to organic mercury, predominantly MeHg, from consumption of fish (Björnberg et al., 2003; Sanzo et al., 2001; Svensson et al., 1992). MeHg exposure in utero is associated with adverse health effects, e.g., neurodevelopmental deficits such as IQ and motor function deficits in children (Mergler et al., 2007; National Research Council, 2000).

In October 2021, the EPA and FDA issued *Advice about Eating Fish and Shellfish* (EPA_FDA, 2021), which updated the consumer advice on mercury in fish originally issued in 2001. This update was due, in part, to research over the past decade that has indicated that fish consumption during pregnancy may be beneficial for the growth and brain development of the fetus and young children (Bramante et al., 2018; Golding et al., 2016; Stratakis et al., 2020; Taylor et al., 2016). The advice provides recommendations for pregnant and breastfeeding women, women of childbearing age, and young children, and includes a chart with fish species considered "Best Choices," "Good Choices," and "Choices to Avoid," with recommended servings per week.

In the <u>FY 2022-2026 EPA Strategic Plan</u>, the EPA has committed in Goal 5 to "Ensure Clean and Safe Water for All Communities" and specifically under Objective 5.2 to "Protect and Restore Waterbodies and Watersheds, to address sources of water pollution and ensure water quality standards are protective of health and needs of all people and ecosystems." The EPA has identified several strategies it will undertake to help protect public health and ensure clean and safe waters that include developing nationally recommended water quality criteria and addressing contaminants that endanger human health. The EPA's approach to making fish safe to eat, which is a human health benefit, has been to:

- Work collaboratively with air agencies to maintain and improve the nation's air quality.
- Encourage development of statewide mercury reduction strategies.

- Reduce air deposition of mercury.
- Improve public information and notification of fish contamination risks.

The Agency can assess progress towards this goal through the measurement of blood mercury concentrations among women of childbearing age as reported by the Centers for Disease Control and Prevention's (CDC) National Health and Nutrition Examination Survey (NHANES). NHANES is a continuous survey designed to collect data on the health and nutritional status of the U.S. population. The NHANES reports data on chemicals, or their metabolites as measured in blood and urine samples collected from a statistically representative sample of the U.S. population. CDC releases NHANES data every two years and reports environmental exposure results for every NHANES release in the National Report on Human Exposure to Environmental Chemicals (Center for Disease Control and Prevention, 2021a, b).

A 2013 study on blood mercury trends in women of childbearing age from NHANES 1999-2010 (Birch et al., 2014; United States Environmental Protection Agency, 2013) found blood mercury concentrations in NHANES survey release 1999-2000 to be statistically significantly higher than the mean of the subsequent 10 releases (2001-2010) for both blood THg and blood MeHg. The EPA reference dose (RfD) for MeHg is 0.1 µg/kg-day (United States Environmental Protection Agency, 2001). This is equivalent to a blood mercury concentration of 5.8 μ g/L. An RfD is an estimate of the maximum daily intake that is not likely to cause harmful effects across a lifetime. From 2008 to 2018, the EPA used the percent of women of childbearing age that have blood mercury concentrations over 5.8 μ g/L as one measure of the progress towards making fish and shellfish safer to eat. In the study of NHANES 1999-2010, the percentage of women of childbearing age with blood MeHg >5.8 μ g/L was significantly higher in survey release 1999-2000. The study also found a significant quadratic trend¹ in blood MeHg concentration since 1999-2000. This quadratic trend indicates decreasing blood MeHg concentrations between NHANES survey release 2001-2002 and 2003-2004, followed by relatively small changes and a slight increase in the survey release 2009-2010. There was a significant relationship between mercury intake from fish consumption and blood MeHg, although mercury intake did not fully explain the differences observed across the survey releases.

A 2009 study (Mahaffey et al., 2009) that investigated regional and coastal differences in NHANES 1999-2004 blood mercury data found that elevated blood mercury occurred more

¹ A non-linear trend described by a second-order polynomial function.

commonly in women living in coastal areas of the United States and that exposure varied regionally with those residing in the Northeast having the highest blood mercury concentrations followed by the South, West, and Midwest.

This study focuses on NHANES 1999-March 2020 data with the goal to investigate national trends in blood mercury concentrations and fish consumption among women of 16-49 years of age. The specific objectives are to assess:

- 1. Trends in blood mercury concentrations over time and by demographic characteristics.
- 2. Association between blood mercury concentration and demographic characteristics.
- 3. Association between estimated usual intake of mercury and blood mercury concentration.
- 4. Trends in frequency and amounts of fish consumed and the association of fish consumption and mercury intake with demographic characteristics.
- 5. Geographic distribution of blood MeHg.

DATA AND METHODS 2

2.1 NHANES and Methods Overview

NHANES is designed to assess the health and nutritional status of adults and children in the United States. It is conducted by the National Center for Health Statistics (Center for Disease Control and Prevention, 2013), part of the CDC that is responsible for producing vital and health statistics for the United States. NHANES collects health-related data from a nationally representative sample of about 5,000 non-institutionalized individuals located in 15 counties in the United States each year and releases the data on two-year cycles. The survey includes interview and examination components. The interview includes demographic, socioeconomic, dietary, and health-related questions. The examination consists of medical, dental, and physiological measurements and laboratory tests of blood and urine. Data from both components of all NHANES releases 1999-March 2020 were used to investigate the trend of blood mercury concentration across NHANES releases by demographic categories. This analysis focused on data from the three most recent NHANES survey releases (2013-2014, 2015-2016, and 2017-March 2020²) to assess fish consumption and compare results with the NHANES 1999-2010 study on trends in fish consumption (Birch et al., 2014; United States Environmental Protection Agency, 2013).

2.2 NHANES Data

The required NHANES data files and variables were identified and downloaded from the NHANES website. These files were merged to create a dataset customized to the needs of this project. For each NHANES survey release, the study data include:

- **Demographics:** Characteristics previously shown to be related to blood mercury and/or fish consumption (gender, age, race/ethnicity, education, and annual income), and sampling weights, pseudo-stratum, and pseudo-primary sampling unit (PSU) variables. The pseudo-stratum and pseudo-PSU variables provide information on how participants were selected and are needed to calculate standard errors and p-values. They are modified from the actual NHANES strata and PSUs for disclosure control and are thus prefixed "pseudo."
- **Survey weights:** Appropriate survey weights of each NHANES release are selected for this analysis (<u>Table 1</u>).

² Due to the COVID-19 pandemic, field operation and data collection for NHANES 2019-2020 were suspended in March 2020. The 2019 March 2020 data were combined with the full data set from NHANES 2017-2018 to create a nationally representative 2017-March 2020 data file.

| Table 1. | Survey weights and ac | Survey weights and adjustment factor for combining multiple NHANES cycles | | | | | | |
|----------|-----------------------|---|------------------------|--|--|--|--|--|
| | Survey release | Survey weight | Adjusted weight factor | | | | | |
| | 1999-2000 | WtMec4Yr*2 | 2/21.2 | | | | | |
| | 2001-2002 | WtMec4Yr*2 | 2/21.2 | | | | | |
| | 2003-2004 | WtMEC2Yr | 2/21.2 | | | | | |
| | 2005-2006 | WtMEC2Yr | 2/21.2 | | | | | |
| | 2007-2008 | WtMEC2Yr | 2/21.2 | | | | | |
| | 2009-2010 | WtMEC2Yr | 2/21.2 | | | | | |
| | 2011-2012 | WtMEC2Yr | 2/21.2 | | | | | |
| | 2013-2014 | WTSH2YR | 2/21.2 | | | | | |
| | 2015-2016 | WTSH2YR | 2/21.2 | | | | | |
| | 2017-March 2020 | WTMECPRP | 3.2/21.2 | | | | | |

Data processing and analyses were performed using the Statistical Analysis System (SAS) version 9.4 (SAS Institute, 2016). In general, two-year Mobile Examination Center (MEC) exam weights (WTMEC2YR) were used for this analysis. The two-year MEC exam weights for NHANES 1999-2000 and 2001-2002 were based on a different population and hence not comparable. An adjusted four-year MEC survey weight (WTMEC4YR) was created to account for the two different reference populations (Johnson et al., 2013).

In the NHANES 2013-2014 and 2015-2016, blood mercury was analyzed in one-half of the 12 years and older population.³ Special sample weights (WTSH2YR) were created for the subsample to account for the additional probability of selection into the subsample, as well as the additional nonresponse to the lab tests results.

- MEC exam weights (WTMECPRP) created for combined NHANES 2017-2018 and 2019-March 2020 data were used for NHANES 2017-March 2020 data.
- In addition, the NHANES 2017-March 2020 data covered 3.2 years compared with 2 years for other NHANES data files. Because the period differs from earlier cycles, the survey weights were adjusted when 2017-March 2020 data files were combined with other two-year cycles (Akinbami et al., 2022). In this study, data from NHANES cycles from 1999-2000 through 2017-March 2020 covered 21.2 years of data collection. For each two-year cycle, the adjusted factor was 2/21.2. And the survey weights for 2017-March 2020 data were adjusted using a factor of 3.2/21.2.
- Laboratory results:
 - Blood total mercury (THg) and inorganic mercury (IHg) concentrations.
 - Hematocrit values, related to blood mercury in that mercury binds to the red blood cells.
- Body measures: Body weight is related to fish consumption and blood mercury.
- Dietary intake, 24-hour recall: Data necessary to estimate usual intake of fish (food codes, meal name, amount eaten; one record per food item eaten). Usual fish intake is

³ NHANES collected data for individuals aged 12 and older. This study uses a subset of the data for women aged 16 to 49.

the long-term average intake of raw finfish and shellfish (from marine, estuarine, and fresh waters).

• Dietary intake, 30 day frequency of consumption: Data used to estimate usual intake of fish (number of times participants reported consuming fish in previous 30 days, calculated from reports for the following species as collected by NHANES – clams, crabs, crayfish, lobster, mussels, oysters, scallops, shrimp, other shellfish, other unknown shellfish, breaded fish products, tuna, bass, catfish, cod, flatfish, haddock, mackerel, perch, pike, pollock, porgy, salmon, sardines, sea bass, shark, swordfish, trout, walleye, other fish, and other unknown fish).

2.2.1 Blood Mercury Data

The laboratory data files from all survey releases contain measurements of THg and IHg in blood. Laboratory methods are available on the <u>NHANES website</u>. <u>Table 2</u> presents the total sample size (N) by survey release, along with the detection limit (DL), the percentage of concentrations observed below the DL, and the standard error of the percentage for blood THg and IHg.

| Cumient | | | THg | IHg | | | | | |
|-------------------|--------|-------------------|--|-----|--------|-------|--|-----|--|
| Survey release | N | DLª | Weighted percent <dl<sup>a</dl<sup> | SE | N | DLª | Weighted percent <dl<sup>a</dl<sup> | SE | |
| All Years | 15,236 | | 14.0 | 0.5 | 15,175 | | 79.6 | 0.7 | |
| 1999-2000 | 1,632 | 0.137 | 5.5 | 0.6 | 1,630 | 0.446 | 97.7 | 0.5 | |
| 2001-2002 | 1,799 | 0.137 | 3.5 | 0.6 | 1,777 | 0.446 | 93.4 | 1.2 | |
| 2003-2004 | 1,615 | 0.20 ^b | 7.6 | 1.3 | 1,595 | 0.446 | 70.2 | 1.9 | |
| 2005-2006 | 1,788 | 0.33 ^b | 15.8 | 1.2 | 1,782 | 0.446 | 64.9 | 3.0 | |
| 2007-2008 | 1,486 | С | 16.7 | 1.6 | 1,484 | 0.446 | 69.4 | 2.3 | |
| 2009-2010 | 1,780 | 0.33 | 15.0 | 1.9 | 1,778 | 0.35 | 84.3 | 1.3 | |
| 2011-2012 | 1,428 | 0.16 | 5.2 | 1.1 | 1,425 | 0.27 | 75.0 | 2.3 | |
| 2013-2014 | 814 | 0.28 | 19.8 | 2.1 | 814 | 0.27 | 79.4 | 2.3 | |
| 2015-2016 | 740 | 0.28 | 17.2 | 1.6 | 739 | 0.27 | 84.0 | 2.4 | |
| 2017-2020 | 2,154 | 0.28 | 25.9 | 2.0 | 2,151 | 0.21 | 78.9 | 1.4 | |

| Table 2. | Sample size and weighted percent of results below the limit of detection, by survey release | |
|----------|---|--|
|----------|---|--|

^a Detection limit. Laboratory Method procedure documents by NHANES releases are available on the NHANES website (<u>https://wwwn.cdc.gov/nchs/nhanes/default.aspx</u>).

^b From Calwell, K.L, et al., 2009

^c Not specified (3*std of >20 runs of blood blank)

2.2.2 Dietary Recall Data

NHANES conducts two 24-hour dietary recalls with participants. The first is an in-person interview and occurs during the examination and the second is a telephone interview 3 to 10 days later. Participants report every food and drink item they consumed in the 24 hours before the interview, including the amount. The food items are coded using U.S. Department of Agriculture (USDA) food codes from the Food and Nutrient Database for Dietary Studies (FNDDS). The FNDDS files are available from the Agriculture Research Service of the USDA (<u>FNDDS DOWNLOAD DATABASES: USDA ARS</u>). The recipes for the food codes were searched to find all food codes that contain finfish or shellfish, including mixed dishes. All records in the 24-hour data file for women aged 16-49 years that were for fish-containing food codes were extracted. The recipe file and 24-hour recall data were merged to calculate quantity of raw fish consumed per recipe. A detailed description of how the fish-containing records were extracted can be found in the 2014 U.S. EPA report, *Estimated Fish Consumption Rates for the U.S. Population and Selected Subpopulations (NHANES 2003-2010)* (United States Environmental Protection Agency, 2014).

Additionally, participants are asked to report the number of times they consumed various types of finfish and shellfish over the past 30 days. Responses to these questions are combined to derive a variable that provides the number of times each respondent consumes any seafood in the past 30 days (frequency of seafood consumption).

2.3 Fish Tissue Mercury Data

To estimate mercury intake, data on mercury concentrations in fish tissue are needed. Mercury concentration in fish varies greatly among species and within species, with older and larger fish having higher concentrations (United States Environmental Protection Agency, 1997). We updated our database of fish tissue mercury concentrations used in the report *Trends in Blood Mercury Concentrations and Fish Consumption Among U.S. Women of Childbearing Age NHANES*, 1999-2010 (United States Environmental Protection Agency, 2013) with data through 2020. This addition included data from the peer reviewed articles (Imanse *et al.*, 2022; Janssen *et al.*, 2021; Janssen *et al.*, 2019; Malinowski, 2019; Melnyk *et al.*, 2021; Sackett *et al.*, 2017; Taylor and Calabrese, 2018; Whitney, 2021; Wolff *et al.*, 2016). Fish tissue mercury concentrations for fish caught between 2011 and 2020 were also extracted from the following databases and reports provided by federal and state governments:

• 2013-2014 National Rivers and Streams Assessment Fish Tissue Study (United States Environmental Protection Agency, 2023).

- Assessment of mercury sources in Alaskan lake food webs: U.S. Geological Survey data release (Lepak, 2022).
- Total Mercury Concentrations in Smallmouth Bass from Chesapeake Bay Tributaries, USA Dataset, 2013-2017 (Willacker, 2020).
- Fish mercury concentration data and ancillary data for streams and rivers across New York States (United States), 1969-2016, including environmental characteristics of selected locations sampled during 2007-16 (Murray, 2020).
- Hg Concentrations of Fish Tissue Samples in the Vicinity of Yellow Pine, Idaho (McGee, 2020).
- Mercury Contaminant Levels in Louisiana Biota (Louisiana Department of Environmental Quality, 2016-2020).
- Selenium and mercury in the Kootenai River, Montana and Idaho, 2018-2019: U.S. Geological Survey data release (Mebane, 2019).
- Status and trends of mercury in fish tissue in New Hampshire waterbodies, 1992-2016 (New Hampshire Department of Environmental Services, 2018).
- Measuring Mercury Trends in Freshwater Fish in Washington State (Washington State Department of Ecology, 2011-2016).
- 2015 Great Lakes Human Health Fish Fillet Tissue Study (United States Environmental Protection Agency, 2015).

Additional data included fish tissue mercury concentrations for the following species that were used for the fish tissue mercury database developed for this analysis: bass, perch, mackerel, carp, catfish, haddock, trout, rockfish, flatfish, sea bass, herring, pompano, pike, salmon, porgy, and whitefish.

To estimate the geometric mean mercury concentration for each fish species, we used the SAS MIXED procedure and modeled the log-transformed fish tissue mercury concentration by fish species, treating the data source as a random effect. Some of the data sources reported average concentrations for multiple fish samples and some sources reported mercury concentrations for each individual fish sampled. In order to account for this in the model, we included a weighting factor. The weighting factor allowed the model to account for different variances due to both data source and number of individual fish samples contributing to each reported value, which modeled the error variance as a power function of the number of samples averaged to obtain the reported value. The predicted values were converted to geometric mean fish mercury concentrations. The average mercury concentration weighted by 30-day consumption frequency was used for fish not specified in the dietary recall food recipe code. To the extent it could be tested, there were no consistent time trend in the fish mercury concentration data in the sources that we used. <u>Table A.1</u> provides the microgram (µg) of mercury per gram of fish by species group used in the analysis.

2.4 Statistical Methods

The relationship between blood mercury concentrations, mercury intake, time, and

demographic characteristics were assessed using:

- 1. Summary statistics based on the imputed concentrations of blood MeHg and the percentage of participants with blood THg and blood MeHg concentrations over 5.8 μ g/L.
- 2. Comparison of imputed concentration of blood MeHg over time by demographic characteristics.
- 3. Estimation of usual intake of fish (Tooze *et al.*, 2010; Tooze *et al.*, 2006; United States Environmental Protection Agency, 2014).
- 4. Regression calibration and linear regression modeling to predict imputed blood MeHg concentration from age, race, education, income, usual intake of mercury, and log-transformed hematocrit concentration (Kipnis *et al.*, 2009).
- 5. Regression modeling to predict amount of fish consumption.
- 6. Logistic regression to predict the probability of reporting any fish consumption in the previous 30 days.
- 7. Regression modeling to predict mercury intake.

For the analysis, we imputed blood MeHg concentration from blood total and inorganic measurements. Blood THg and IHg measurements below the DL were also imputed. Details of the estimation of usual intake of mercury and the methodology to model MeHg concentrations are described in Section 2.4.1. Details of the imputation methodology are discussed in Section 2.4.2.

A box plot and table of blood MeHg concentrations were generated to provide sample sizes, geometric means, 95 percent confidence intervals of geometric means, and percentiles (25th, 75th, and 90th) by survey release. Geometric means of blood MeHg concentrations and their standard errors were generated by age group, race/ethnicity, income, and education. A test with p value less than 0.05 is considered significant. These data were generated by averaging the 20 imputed values for each individual, then calculating the statistics from those values utilizing a SAS software survey procedure to incorporate the uncertainty due to the survey sample design. Data from all NHANES survey releases were used to investigate the trend of blood MeHg concentration over time (i.e., 1999-2004, 2005-2010, 2011-2016, and 2017-March 2020) (Table 4).

The frequency distribution of fish consumption, estimated 30-day fish consumption amount, estimated 30-day mercury intake, and the estimated 30-day mercury intake per unit body weight were calculated. Detailed tables of these distributions were generated to provide sample sizes, arithmetic means, percentiles (25th, 50th, 75th, 90th, and 95th) and their 95 percent confidence intervals, by survey releases, race/ethnicity, age, income, and education. These tables are in the Appendix (Tables <u>A.4a</u>, <u>A.5</u> to <u>A.8</u>). Plots and analytic extracts based on these tables are presented in the report.

Data processing and analyses were performed using SAS software, 9.4 (SAS Institute, 2016) and following the NHANES Analytical Guidelines posted on the NHANES website (Akinbami *et al.*, 2022). All analyses were weighted using the statistical weights recommended in the guidelines as detailed in Section 2.2, and the sampling design variables were used in calculating the variance of the estimates.

Age was categorized into four groups: 16-19, 20-29, 30-39, and 40-49 years. Race/ethnicity groups recorded in NHANES consistently across the 10 survey periods include Mexican American, Other Hispanic, Non-Hispanic White, Non-Hispanic Black, and "Other Race." "Other Race" consists of Asian, Native American, Pacific and Caribbean Islander, Alaska Native, multiracial, and unknown race.

Hematocrit was included in the analysis because approximately 80 percent of MeHg binds to red blood cells (Clarkson and Magos, 2006; Rothenberg *et al.*, 2015). This variable was logtransformed for modeling.

Some demographic information (e.g., family and household income; education for youth 6-19 years) were modified or not included in the 2017–March 2020 public-use data file release due to potential disclosure risks (Akinbami *et al.*, 2022).

Family and household income is not included in the 2017–March 2020 release. The ratio of family income to the federal poverty level is included as for previous cycles. The seven income categories used for the analysis are based on this ratio: less than one times the ratio, one to less than two times the ratio, two to less than three times the ratio, three to less than four times the ratio, four to less than five times the ratio, greater than or equal to five times the ratio, and missing values.

Education level for adults aged 20 and over in the 2017-March 2020 release is included. Education is categorized as less than, equal to, or greater than the median education level for the participant's age for all previous NHANES cycles, and adults aged 20 and over for NHANES 2017-March 2020. Education level for children and youth of 6–19 years was not included in the 2017–March 2020 release and thus for this study, participants aged 16-19 years are categorized as an unknown education group.

2.4.1 Usual Intake of Fish and Mercury

The National Cancer Institute (NCI) provides SAS macros (titled MIXTRAN and INDIVINT) to calculate the distribution of usual intake of dietary components (such as fish and mercury from fish) and for calculating the expected mean of transformed usual intake (TUI) that can be used for calculating the relationship between usual intake and another dependent variable, such as blood mercury concentrations. Using fish consumption as an example, the MIXTRAN macro fits the following models in order to predict the distribution of usual intake across the population:

- A logistic mixed model for predicting the probability of consuming fish in any 24-hour dietary recall.
- A linear mixed model to predict the reported amount of fish consumed in dietary recalls where fish consumption is reported. The fish consumption is modeled on a transformed scale using a Box-Cox transformation.
- The usual intake is the product of the probability of fish consumption and the amount consumed.

The output from the MIXTRAN macro can be used as input to the INDIVINT macro to calculate the expected mean fish consumption for each individual or equivalently the regression calibration estimate of fish consumption that can then be used to predict blood mercury concentration in a linear regression model.

The NCI macros have some limitations when applied to the NHANES fish consumption data, including:

- With many predictors, the MIXTRAN macro may take a long time to converge to the final parameter estimates or may fail to converge.
- When assuming the random effects in the logistic and linear mixed models are correlated, MIXTRAN may fail to converge when using the NHANES survey weights.
- With the NHANES data, the MIXTRAN macro needs to be run multiple times, once for each replicate weight created to calculate variance of the parameter estimates.

Due to these limitations, Westat created a modification to the NCI method, hereafter referred to as the EPA method. The EPA method has the following steps:

- 1. Fit the weighted logistic model using the SAS SURVEYLOGISTIC procedure, saving the logit transformed predicted probability of fish consumption (call this LogitXBeta).
- 2. To estimate the variance components, fit a logistic mixed model similar to that used by MIXTRAN except the only predictor is LogitXBeta and the intercept is set to zero.
- 3. Determine a Box-Cox transformation that makes the reported distribution of amount consumed roughly normally distributed.

- 4. Fit a weighted linear model to predict the amount of fish consumed, when consumed, saving the predicted amount.
- 5. To estimate the variance components, fit a linear mixed model similar to that used by MIXTRAN, except the only predictor is the predicted amount from the previous regression and the intercept and slope are set to one and zero, respectively.
- 6. Assuming random effects from the logistic and linear models are independent, the output from previous modeling steps are used as inputs to the NCI INDIVINT macro to calculate the regression calibration estimate for transformed usual fish consumption. See the discussion below regarding selection of the transformation to use.
- 7. Finally, the regression calibration estimate of usual fish consumption is used to predict blood mercury concentration using weighted regression.

The steps above were repeated for each replicate weight.

Regression calibration assumes that a linear model is used to predict the dependent variable (blood mercury concentration) from the transformed usual intake. The INDIVINT macro assumes a Box-Cox transformation of usual intake is used and requires that a Box-Cox lambda parameter be specified. Several values of lambda were tried to identify a lambda for which the relationship between the regression calibration estimates of usual intake were most linearly related to blood THg measurements. A Box-Cox lambda of 0.70 was selected and used for all analyses. Because a value of 0.70 might not be optimal for all analyses, the final regression model allowed for the slope above and below the median usual intake to differ. In all cases the slope difference was not statistically significant. To accommodate this modification, non-linear regression was used to fit the final model.

2.4.2 Imputation

The MeHg concentration was calculated from the difference between the THg measurements and the IHg measurement. However, due to measurement errors, that difference can be negative. In addition, a few of the THg concentrations were less than the DL and not otherwise specified and many of the IHg measurements were below the DL.

Predictors were selected by stepwise selection using SAS GLMSELSECT to predict the log-transformed blood THg concentration from calculated TUI and selected main effects. Significant (p<0.05) main effects and interactions of (TUI) and main effects were identified. The selected predictors were then tested for significance using SAS SURVEYREG. The final set of main effect variables were race/ethnicity, income, education, NHANES survey release, centered age in decade, centered log-transformed body weight, centered log-transformed hematocrit, and TUI. The final set of two-way interactions were interactions of TUI with

centered log-transformed body weight, centered log-transformed hematocrit, race/ethnicity, and education.

The following procedure was used to impute the missing concentrations for the non-detects and adjust for the negative values. A Bayesian model was used to:

- Impute the THg values less than the DL as a function of the regression calibration estimate of usual intake (using the NHANES analysis weight) and other predictors.
- Impute the IHg values less than the DL as a function of the regression calibration estimate of usual intake (using the NHANES analysis weight), the THg measurement (detected or imputed) and other predictors.
- Calculate the preliminary organic mercury concentration as the difference between the total and inorganic mercury concentrations (using detected or imputed values).
- Adjust the differences to be greater than zero such that the log-transformed differences have a roughly normal distribution.

The following transformation was used to adjust the smallest differences upward to make all values greater than zero.

$$Organic Hg = \frac{Difference + \sqrt{Difference^2 + 0.04}}{2}$$

Twenty imputed datasets were created for the analysis. For each replicate weight, the final regression model (Step 7 in the EPA Method) was fit separately for each of the 20 imputed datasets.

2.5 Estimation of 30-Day Fish Consumption and Mercury Intake

In order to investigate the relationship between fish consumption and mercury intake with demographic characteristics, estimates of the amount of fish consumed over 30 days were calculated based on the NHANES 24-hour dietary recall data and the 30-day frequency of fish consumption data. The 24-hour data provided the amount of meal consumed at one time. Information on the amount and species of fish ingredients provided in the 24-hour recall can be obtained by linking the FNDDS database with food code. Many participants who reported consuming fish during the previous 30 days did not consume fish in the past 24-hours, so a corresponding amount and species of fish consumed in a meal were not available. Therefore, a statistical model was used to estimate the amount of fish they consumed in a meal to calculate their estimated 30-day consumption. We followed the method described in the report *Trends in Blood Mercury Concentrations and Fish Consumption Among U.S. Women of Childbearing Age NHANES*, 1999-2010 (United States Environmental Protection Agency,

2013) to estimate the amount of 30-day consumption of fish. The predicted grams of fish consumed in a meal for each species from the model were multiplied by the reported frequency of consumption of corresponding species to get the amount of 30-day consumption for each fish species. The sum of the 31 species-specific fish consumption amounts of each participant was the estimated 30-day consumption of fish for each women aged 16-49 years. The estimated mercury intake was calculated as the product of species-specific fish tissue mercury concentration and the estimated amount of fish consumed at one time for each species. The participant level 30-day mercury intake was calculated as the sum of mercury intake was calculated as the sum of fish consumed at one time for each species.

Mercury intake per body weight can be explained as the product of four components: 1) frequency of fish consumption; 2) weighted average meal size, weighted by frequency of consumption; 3) weighted average fish tissue mercury concentration, weighted by the quantity of fish consumed; and 4) inverse body weight (United States Environmental Protection Agency, 2013).

Logistic regression was used to model the probability of consuming fish in a 30-day period. For participants who reported consuming fish, regression analysis was used to investigate the association between these four components and demographic characteristics (education, race/ethnicity, income, and age group).



3.1 Blood MeHg Summary Statistics

This section presents the summary statistics based on the imputed concentrations of blood MeHg by selected subpopulations.

3.1.1 Time Trends in Blood Mercury Concentrations

Figure 1 presents the distribution of blood MeHg concentration by NHANES survey release using boxplots. The geometric mean blood MeHg concentration is highest in survey release 1999-2000 then declines to the lowest in 2017-March 2020. The geometric mean blood MeHg in the 2005-2006 release slightly higher than that in the 2003-2004 and 2007-2008 releases, and the geometric mean blood MeHg in the 2009-2010 release slightly higher than that in the 2007-2008 and 2011-2012 releases. This general declining trend in geometric means is significant (p<0.001) across the ten survey releases. This same general pattern over time is observed in the 25th and 90th percentiles of blood MeHg concentrations (<u>Table 3</u>). Detailed tabulations of the distribution of blood MeHg concentrations, including sample size, arithmetic mean, geometric mean, percentiles (25th, 50th, 75th, 90th, 95th), and their 95 percent confidence intervals, by NHANES releases, race/ethnicity, age group, income, and education are presented in <u>Table A.2</u>.

Figure 1. Distribution of log-transformed blood MeHg (µg/L), by NHANES survey releases, women aged 16-49 years

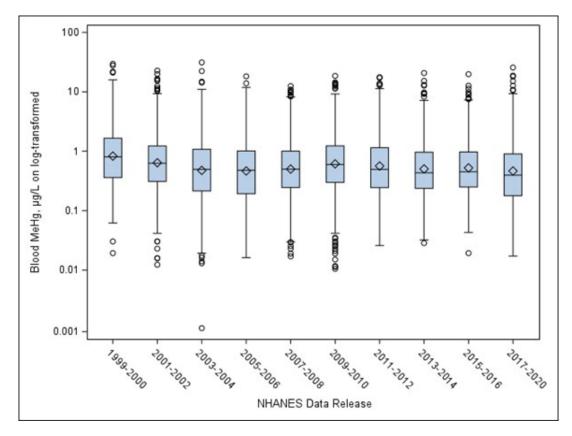


Table 3. Distribution of blood MeHg concentration (µg/L), by survey release

| Company and a mage | N | Geometric mean | | Percentiles | | | |
|--------------------|--------|-------------------|------|-------------|------|--|--|
| Survey release | N - | (95% CI) | 25th | 75th | 90th | | |
| All Years | 15,236 | 0.57 (0.55, 0.60) | 0.25 | 1.21 | 2.58 | | |
| 1999-2000 | 1,632 | 0.93 (0.78, 1.12) | 0.41 | 1.88 | 4.56 | | |
| 2001-2002 | 1,799 | 0.72 (0.65, 0.79) | 0.34 | 1.46 | 2.92 | | |
| 2003-2004 | 1,615 | 0.53 (0.45, 0.61) | 0.24 | 1.21 | 2.60 | | |
| 2005-2006 | 1,788 | 0.58 (0.51, 0.66) | 0.25 | 1.33 | 2.71 | | |
| 2007-2008 | 1,486 | 0.52 (0.46, 0.59) | 0.25 | 1.06 | 2.39 | | |
| 2009-2010 | 1,780 | 0.67 (0.61, 0.73) | 0.31 | 1.38 | 2.73 | | |
| 2011-2012 | 1,428 | 0.52 (0.44, 0.60) | 0.23 | 0.98 | 2.32 | | |
| 2013-2014 | 814 | 0.50 (0.45, 0.55) | 0.23 | 0.94 | 2.21 | | |
| 2015-2016 | 740 | 0.50 (0.45, 0.56) | 0.24 | 0.94 | 2.19 | | |
| 2017-2020 | 2,154 | 0.48 (0.43, 0.54) | 0.18 | 1.01 | 1.99 | | |

The EPA reference dose (RfD) for MeHg is 0.1 μ g/kg-day (United States Environmental Protection Agency, 2001). This is equivalent to a blood mercury concentration of 5.8 μ g/L. An RfD is an estimate of the maximum daily intake that is not likely to cause harmful effects across a lifetime. The EPA uses the percent of women of childbearing age that have blood mercury concentrations over 5.8 μ g/L as one measure of the progress towards making fish and shellfish safer to eat. The calculated weighted prevalence of both blood MeHg and THg concentrations over 5.8 μ g/L by survey release are presented in <u>Table 4</u>. The percentages of women of childbearing age with blood MeHg concentrations above 5.8 μ g/L in 1999-2000 is about 3.5 times that found in 2001-March 2020, representing a 71 percent decrease. There is a significant difference between the survey releases for MeHg (Rao-Scott Chi-square p<0.001). No significance difference (p=0.12) is found between the survey releases after removing the 1999-2000 survey release.

Blood MeHg Blood THg Survey release Ν Weighted % SE Weighted % SE 1,632 1.65 1.67 1999-2000 6.9 7.3 2001-2002 1,799 3.5 0.83 3.9 0.88 2003-2004 1,615 1.7 0.69 2.5 0.83 2005-2006 0.58 2.7 0.60 1,788 2.4 2007-2008 1,486 2.3 0.50 2.5 0.55 2009-2010 1,780 2.1 0.37 2.3 0.41 2011-2012 1,428 1.3 0.45 1.8 0.50 2013-2014 814 2.3 0.69 2.7 0.66 2015-2016 740 1.8 0.50 1.8 0.49 2017-2020 2,154 1.2 0.37 1.3 0.37

Table 4.Weighted percent of women 16 and 49 years with blood MeHg \geq 5.8 µg/L and blood THg \geq 5.8
µg/L, by survey release

Note: Geometric mean and percentiles were calculated from the mean of imputed values for each respondent. Decreasing trend over time, p<0.001 for both blood MeHg and THg concentrations based on logistic regression.

3.1.2 Demographic Distributions

Table 5 presents the comparison of blood MeHg concentration over time by demographic characteristics for women 16-49. NHANES survey periods are grouped as 1999-2004, 2005-2010, 2011-2016, and 2017-March 2020. The geometric mean of blood MeHg concentrations decreases significantly over time in most demographic categories. There is a 31 percent decrease in blood MeHg concentration from 1999-2004 to 2017-March 2020. Within each time period, people of "Other Race," 40-49 years of age, with income greater than or equal to 5 times the poverty line and education level higher than median level of the age group, have higher blood MeHg concentrations compared to the rest of the groups within each demographic characteristic.

| Characteristics | , | agea lo | 10 9 0010 | | | | | | |
|---|--------------|---------|--------------|------|--------------|------|--------------------|------|----------|
| Blood mercury | 1999- | 2004 | 2005-2010 | | 2011-2016 | | 2017-March 2020 | | p-value |
| concentrations | Geo. mean | SE | Geo. mean | SE | Geo. mean | SE | Geo. mean | SE | (F test) |
| OVERALL | 0.70 | 0.03 | 0.59 | 0.02 | 0.51 | 0.02 | 0.48 | 0.03 | <.0001 |
| Race/Ethnicity | | | | | | | | | |
| Mexican American | 0.53 | 0.03 | 0.44 | 0.02 | 0.39 | 0.02 | 0.42 | 0.03 | 0.0027 |
| Other Hispanic | 0.85 | 0.07 | 0.66 | 0.05 | 0.59 | 0.04 | 0.54 | 0.05 | 0.0320 |
| Non-Hispanic White | 0.66 | 0.04 | 0.55 | 0.03 | 0.47 | 0.02 | 0.45 | 0.05 | 0.0003 |
| Non-Hispanic Black | 0.90 | 0.06 | 0.65 | 0.03 | 0.53 | 0.04 | 0.56 | 0.05 | <.0001 |
| Other Race | 1.17 | 0.13 | 1.25 | 0.12 | 0.89 | 0.07 | 0.64 | 0.05 | <.0001 |
| Age, Years | | | | | | | | | |
| 16 to 19 | 0.44 | 0.03 | 0.37 | 0.02 | 0.31 | 0.02 | 0.33 | 0.03 | <.0001 |
| 20 to 29 | 0.60 | 0.03 | 0.52 | 0.03 | 0.47 | 0.03 | 0.49 | 0.06 | 0.0122 |
| 30 to 39 | 0.80 | 0.06 | 0.63 | 0.04 | 0.55 | 0.03 | 0.51 | 0.03 | <.0001 |
| 40 to 49 | 0.83 | 0.05 | 0.71 | 0.03 | 0.61 | 0.03 | 0.52 | 0.04 | <.0001 |
| Ratio of Family Income to Po | overty Gu | | | | | | | | |
| 0 to <1x poverty line | 0.54 | 0.03 | 0.44 | 0.03 | 0.39 | 0.02 | 0.41 | 0.04 | 0.0006 |
| 1x to <2x poverty | 0.58 | 0.03 | 0.48 | 0.02 | 0.40 | 0.02 | 0.38 | 0.02 | <.0001 |
| 2x to <3x poverty line | 0.68 | 0.05 | 0.53 | 0.04 | 0.49 | 0.04 | 0.41 | 0.03 | 0.0003 |
| 3x to <4x poverty line | 0.65 | 0.06 | 0.57 | 0.04 | 0.53 | 0.05 | 0.49 | 0.06 | 0.1904 |
| 4x to <5x poverty line | 0.77 | 0.05 | 0.60 | 0.05 | 0.54 | 0.04 | 0.52 | 0.07 | 0.0077 |
| >= 5x poverty line | 1.07 | 0.08 | 0.94 | 0.06 | 0.81 | 0.06 | 0.68 | 0.06 | 0.0012 |
| Missing/Refused/DK | 0.78 | 0.09 | 0.63 | 0.07 | 0.60 | 0.04 | 0.54 | 0.07 | 0.2421 |
| Education | | | | | | | | | |
| <median age<="" education="" for="" td=""><td>0.59</td><td>0.03</td><td>0.46</td><td>0.02</td><td>0.39</td><td>0.02</td><td>0.38</td><td>0.03</td><td><.0001</td></median> | 0.59 | 0.03 | 0.46 | 0.02 | 0.39 | 0.02 | 0.38 | 0.03 | <.0001 |
| =Median education for age | 0.67 | 0.03 | 0.57 | 0.03 | 0.46 | 0.02 | 0.47 | 0.03 | <.0001 |
| >Median education for age | 0.98 | 0.07 | 0.82 | 0.05 | 0.72 | 0.04 | 0.69 | 0.07 | 0.0039 |

Table 5.Comparison of imputed blood MeHg concentration (µg/L) over time, by demographic
characteristics, women aged 16-49 years

3.2 Blood MeHg Modeling

The parameter estimates and p-values from the multivariable modeling of blood MeHg concentrations are presented in <u>Table A.3</u>. The parameter estimate provides the direction and magnitude of the effect the predictor has on log-transformed blood MeHg concentration. Level of education (p=0.0007) and race/ethnicity (p<0.001) are significantly associated with log-transformed blood MeHg concentrations. Figure 2 shows the relative blood MeHg concentrations for different education and racial/ethnic groups represented by multiplicative difference from the overall MeHg concentration. The red diamond shows the estimate and the orange bar shows the 95 percent confidence interval of the estimate. If the 95 percent confidence interval does not cross 1, then the estimate is statistically significant at p<0.05. Blood MeHg concentration increases with education level. The difference in blood MeHg concentrations between women with education levels above the median for their age and those at the median level is more pronounced than the difference between women with education levels at the median and those below it for their age. There is significant difference

(p<0.0001) in blood MeHg concentrations between the race/ethnicity groups, with non-Hispanic White women of childbearing age having the highest blood MeHg concentrations and Mexican American having the lowest.

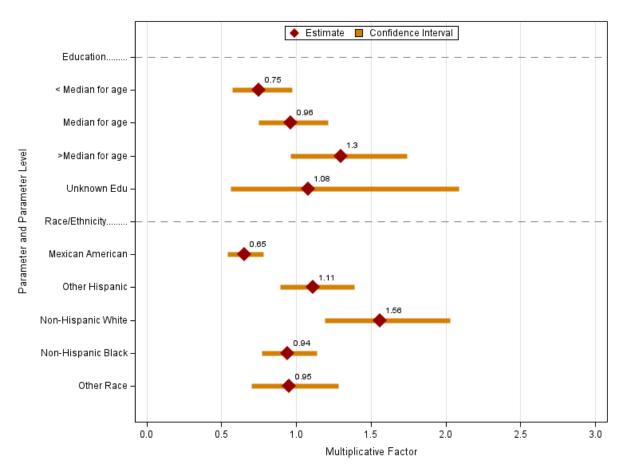


Figure 2.Relative blood MeHg concentrations with 95 percent confidence intervals, by demographic
characteristics (NHANES 2013-March 2020)

TUI of mercury (µg Hg/day) through fish consumption is one of the significant predictors of blood MeHg concentration (p<0.001), with higher consumption associated with higher blood MeHg concentration.

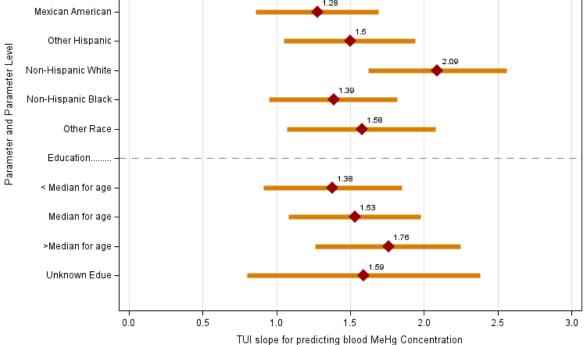
Figure 3 shows the slopes for the relationship between TUI of mercury through fish consumption and log-transformed MeHg overall and by demographic group. The strength of the relationship between usual intake of mercury through fish consumption and the blood MeHg concentration increases as the value of the slope factor increases. The following provides a basis for interoperating the TUI slope parameters. Consider comparing blood MeHg concentrations between subjects whose usual intake of mercury from fish consumption differs by 10 percent while holding all other factors constant. If all MeHg comes from mercury intake through fish intake, increasing usual intake of mercury through fish consumption by 10 percent should increase blood MeHg concentration by 10 percent (TUI slope =0.1). If half of the blood MeHg concentrations comes from usual intake of mercury through fish consumption and half comes from other sources, increasing usual intake of mercury through fish consumption by 10 percent should increase blood MeHg concentrations by 5 percent (TUI = 0.05). The slope relating log-transformed usual intake of mercury through fish consumption to log-transformed MeHg is an estimate of the fraction of blood MeHg concentration from usual intake of mercury through fish consumption. The slope for TUI is a close approximation to the slope of the relationship between logtransformed usual intake and log-transformed blood MeHg concentration.

All slopes are greater than zero, indicating a positive relationship between TUI of mercury through fish consumption and blood MeHg concentration. The slope varies significantly based on level of both education (p=0.01) and race/ethnicity (p<0.001), indicating that blood MeHg concentration increases at different rates with increasing intake of fish mercury by education and race/ethnic groups. Participants whose education level is higher than the median education level of the corresponding age group have a higher slope than the rest of the education groups indicating that blood MeHg concentration increase more to a unit increase in TUI of mercury for those with greater than median level education compared to the other participants. Non-Hispanic White participants have the highest slope among all race/ethnic groups and therefore have blood MeHg concentrations that increase faster as TUI of mercury increases.

blood MeHg concentrations, overall and by demographic group, with 95 percent confidence intervals

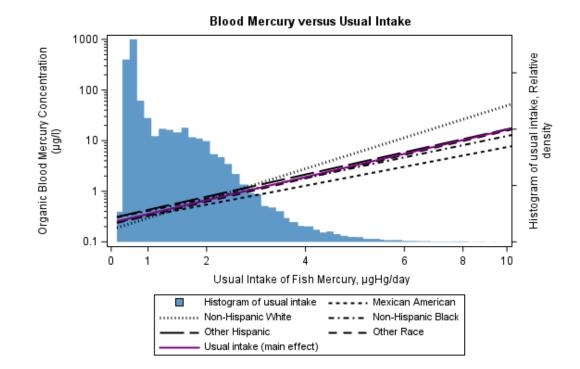
Slope parameter relating transformed usual fish intake of mercury and log-transformed

Figure 3.



Due to the model's complexity, the slope parameters presented in Figure 3 may be difficult to interpret in terms of TUI of mercury (µg Hg/day). Figure 4 shows the predicted relationship between blood MeHg concentrations and TUI of mercury by racial/ethnic groups. For non-Hispanic White participants, blood MeHg concentrations increase at a faster rate with TUI of mercury compared to the rest of the groups. The slope is similar among participants who identify as Mexican American, other Hispanic, and non-Hispanic Black.





Other significant predictors include NHANES survey release (p<0.001), log-transformed body weight (p=0.002) and its interaction with TUI of mercury (p<0.001), and log-transformed hematocrit concentration (p=0.005).

3.3 Trends in Fish Consumption

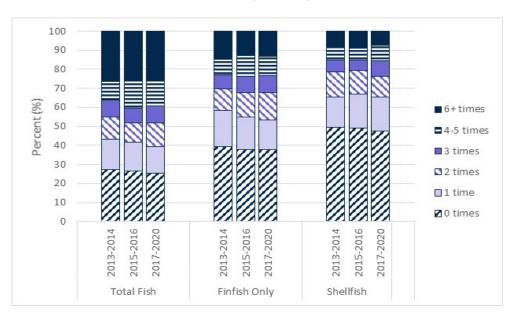
This section presents trends in fish consumption of NHANES 2013-March 2020 and compares them to trends in fish consumption of NHANES 1999-2010 (United States Environmental Protection Agency, 2013). NHANES 2011-2012 data was not included in the current study and therefore not part of this trends analysis.

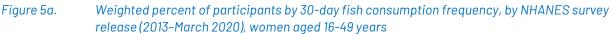
3.3.1 Trends in Frequency of Consumption

Figure 5a presents the weighted percent of women aged 16-49 years in each of the six categories of reported frequency of fish consumption by NHANES survey releases (2013-March 2020). Detailed tabulations are in <u>Table A.4a</u>. There are no significant differences in reported frequency of consumption between survey releases (Rao-Scott chi-square p-values: p=0.66 for total fish, p=0.56 for finfish, p=0.74 for shellfish). Figure 5b displays the same distribution of NHANES 1999-2010 (United States Environmental Protection Agency, 2013). While there are statistically significant differences in consumption frequency between the

survey releases 1999-2010, there is not a consistent trend over time (Rao-Scott chi-square p-values; p=0.03 for total fish, p=0.02 for finfish, and p=0.16 for shellfish). Comparing the frequency of consumption of total fish between the two studies, there are 7 percent more women reported not consuming fish in the previous 30 days in 2013-2014 than in 2009-2010, and approximate 5 percent decrease in women reported consuming fish 6 times or more from 2009-2010 (31.8%) to 2013-2014 (26.4%). This indicates a shift of consuming fish less frequently in women of childbearing age.

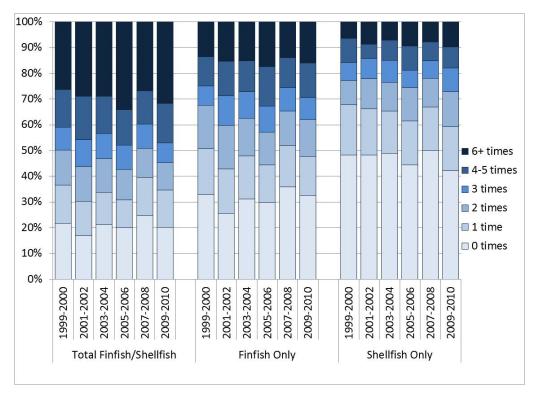
The fish consumption frequencies are similar between survey releases for each consumption category. However, there are some differences in the frequency of consumption by category of fish. For example, in NHANES 2013-2014, the percent of women who did not consume total fish (finfish and shellfish) in the previous 30 days (27%) is less than the percent of women who did not consume finfish (39%) or shellfish only (49%) in the previous 30 days. This is likely due to the fact that some women only consumed finfish or shellfish. Similarly, the percent of women who consumed fish a total of six or more times in the previous 30 days (26%) was greater than the percent of women who consumed finfish (15%) or shellfish (8%) only six or more times. This difference is a result of participants who may have consumed finfish or shellfish less than six times, but when combined, they consumed either finfish or shellfish six or more times. These findings are similar to a previous study of NHANES 1999-2010 data (United States Environmental Protection Agency, 2013).





Note: Data for Figure 5a is in <u>Table A.4a</u> in Appendix A.

Figure 5b. Weighted percent of participants by 30-day fish consumption frequency, by NHANES survey release (1999-2010), women aged 16-49 years



Note: Data for Figure 5b is in <u>Table A.4b</u> in Appendix A.

Figure 6a presents the frequency of consumption by income, race/ethnicity, education, and age of NHANES 2013-March 2020. There are significant differences in frequency of total fish consumption between income groups, race/ethnicity, age, and education (Rao-Scott Chi-Square p-values <0.0001). These demographic characteristics were included in the analysis of relationship between fish consumption and intake of mercury with demographic factors. Women with higher income tend to eat fish more frequently. Individuals of "Other Race" eat fish more frequently compared to Mexican American, non-Hispanic White, non-Hispanic Black, and other Hispanic. These findings are consistent with a previous study of NHANES 1999-2010 data (United States Environmental Protection Agency, 2013) shown in Figure 6b. Older age is associated with increased frequency of fish consumption in both studies. Women aged 30-39 years consume fish more frequently than those in other age groups of NHANES 1999-2010 (United States Environmental Protection Agency, 2013). Women with higher levels of education are associated with higher frequency of fish consumption based on NHANES 2013-March 2020.

There is a trend of decreasing frequency of fish consumption among women of "Other Race," non-Hispanic White, and ages 30-39 and 40-49. For example, comparing the estimates by race/ethnicity between NHANES 1999-2010 and 2013-March 2020, the percent of women of "Other Race" who reported total fish consumption of 6 or more times decreased from 46.6 percent to 35.4 percent, and the reported frequency of not consuming fish increased from 19.5 percent to 22.3 percent. Similar patterns were also observed in Non-Hispanic White.

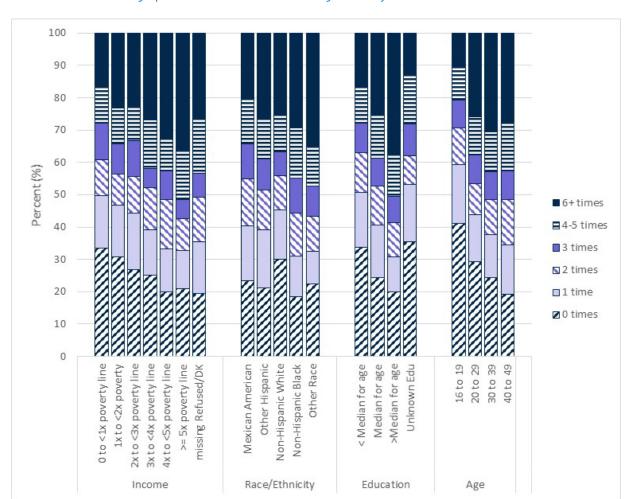


Figure 6a. Weighted percent of participants by 30-day total fish consumption frequency, by demographic characteristics, women aged 16-49 years, NHANES 2013-March 2020

Note: Data for Figure 6a is in Table A.4a in Appendix A.

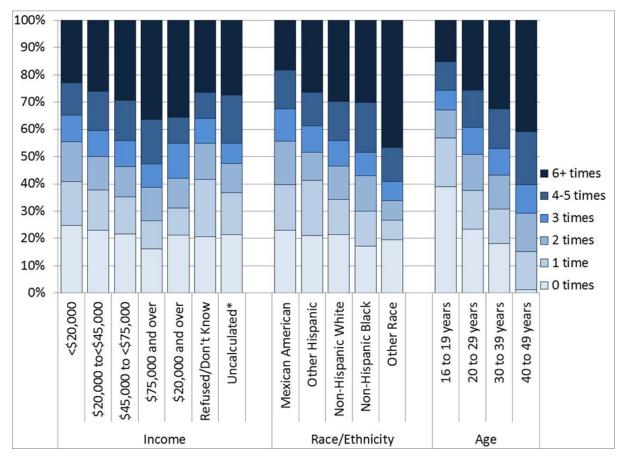


Figure 6b. Weighted percent of participants by 30-day total fish consumption frequency, by demographic characteristics, women aged 16-49 years, NHANES 1999 – 2010

Note: Data for Figure 6b is in <u>Table A.4b</u> in Appendix A.

* Uncalculated indicates that the participant is residing in a multi-family dwelling and one or more of the families only reported a range for their family income, either <\$20,000 or >\$20,000. Thus NCHS did not calculate household income for these participants.

3.3.2 Trends in Estimated Amounts Consumed Over the Previous 30 days

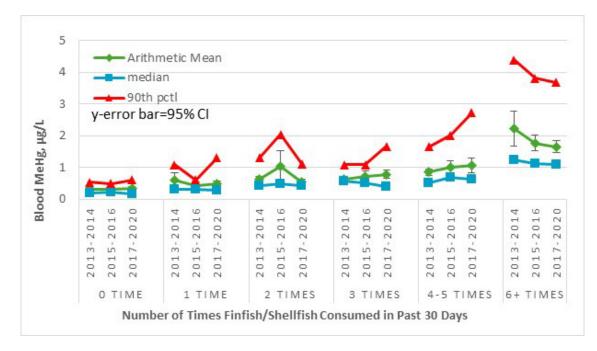
The estimated amounts of fish consumed has remained relatively consistent over the NHANES survey releases from 2013-March 2020. Detailed tabulation of the amounts of fish consumed (in grams [g]), mercury intake (in micrograms [µg]), and mercury intake per unit body weight (µg Hg/kg bw) are tabulated in Tables A.5 through A.7 by NHANES releases, and Table A.8 by race/ethnicity, age, income, and education. While the average amounts of total fish eaten per participant and NHANES release for NHANES 2013-March 2020 (318.8-335.3 g) are at the higher end of those found in NHANES 1999-2010 (254.6-322.5 g) (United States Environmental Protection Agency, 2013), the estimated mercury intake from total fish consumption in NHANES 2013-March 2020 (22.83-25.62 µg) are lower than those of the

previous survey releases (29.33-37.40 μ g), indicating women probably choose to eat fish with lower mercury concentration. This leads to lower estimates of mercury intake per unit body weight in NHANES 2013-March 2020 (0.32-0.36 μ g Hg/kg bw) compared to those found in NHANES 1999-2010 (0.42-0.54 μ g Hg/kg bw).

3.3.3 Association Between Fish Consumption Frequency and Blood Mercury

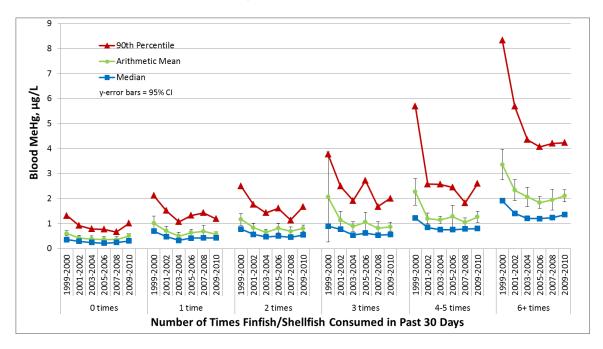
Figure 7a presents the distribution of mean blood MeHg concentrations by the 30-day frequency of total fish consumption and NHANES survey release (2013-March 2020). Detailed tabulations are in Table A.9a. Blood MeHg concentration increases with frequency of fish consumption (p<0.0001). This agrees with previous studies that people who eat fish more frequently tend to have higher blood mercury concentrations (Birch et al., 2014; Mahaffey et al., 2004; Mahaffey et al., 2009; United States Environmental Protection Agency, 2013). The distribution of blood MeHg concentrations over time are not consistent by frequency of consumption groups. The arithmetic mean of blood MeHg concentration for women who ate fish six or more times decreases from 2.22 (1.66, 2.77) µg/L in 2013-2014 to 1.66 (1.47, 1.85) µg/L in 2017-March 2020. For women who ate fish two times, the mean blood MeHg concentration in 2005-2006 (1.04(0.55,1.53) µg/L) is higher than that in survey releases 2013-2014 and 2017-March 2020. The blood MeHg concentrations have small variations for those who do not eat fish or ate fish one, three, four and five times. Based on the study of fish consumption of NHANES 1999-2010 as presented in Figure 7b (United States Environmental Protection Agency, 2013), there is statistically significant decreasing trend of blood MeHg concentration over time, indicating that women who consume fish more often may be shifting to fish with lower concentrations of mercury. Detailed tabulations for Figure 7b are in Table A.9b.

Figure 7a. Mean blood MeHg concentrations by reported frequency of total fish consumption in 30 days, women aged 16-49 years, NHANES 2013-March 2020 (with 95% confidence intervals, median, and 90th percentile)



Note: Data for Figure 7a is in <u>Table A.9a</u> in Appendix A.

Figure 7b. Mean blood MeHg concentrations by reported frequency of total fish consumption in 30 days, women aged 16-49 years, NHANES 1999–March 2010 (with 95% confidence intervals, median, and 90th percentile)



Note: Data for Figure 7b is in <u>Table A.9b</u> in Appendix A.

3.3.4 Relationship Between Fish Consumption and Intake of Mercury with Demographic Factors

This section examines the statistical association between fish consumption and demographic characteristics using the method detailed in <u>Trend in Blood Mercury</u> <u>Concentrations and Fish Consumption Among U.S. Women of Childbearing Age NHANES</u> <u>1999-2010</u> (United States Environmental Protection Agency, 2013). Estimates of the amount of fish consumed in the previous 30 days, mercury intake, and mercury intake per unit body weight were calculated using the method described in section 2.5. Logistic regression was applied to model the probability of a person reporting any fish consumption in the previous 30 days using education, race/ethnicity, income, and age. For those who reported consumption of fish, five regression models were fit to predict fish consumption and mercury intake variables from demographic characteristics. The five variables were (1) mercury intake per unit body weight and the four components of this variable, (2) number of meals in the previous 30 days, (3) amount of fish consumed in a meal, (4) the mercury concentration in the fish consumed calculated as the ratio of mercury intake to fish consumption in the previous 30 days, and (5) the inverse of body weight. The model results are presented in <u>Tables A10</u> to A.15.

Figure 8 presents the results from the logistic regression models by education groups of NHANES 2013-March 2020. Education is categorized as less than, equal to, or greater than the median education level for the participant's age for NHANES releases 2013-2014, 2015-2016, and 2017-March 2020, and an additional unknown level for participants 16-19 years in NHANES 2017-March 2020. The percentages in parentheses next to the education group, e.g., <median for age (30%), are the percent of participants categorized in the education group with lower than median education level for age. The grey star and the error bars plotted on the second y-axis are the percent of women who reported fish consumption in the previous 30 days. The remaining colored symbols and error bars are the relative ratios (RR) and 95 percent confidence intervals from the regression models predicting (1) the log-transformed frequency of fish consumption in the previous 30 days (maroon open diamond); (2) the logtransformed amount fish in a meal (meal size, green filled diamond); (3) the log-transformed mercury concentration in the fish consumed (red filled diamond); (4) the log-transformed inverse of body weight (blue filled dot); and (5) the log-transformed mercury intake per unit body weight (black filled square). The horizontal line on the plot at RR=1 represents the geometric mean response for a hypothetical population equally divided among categories for education. This line is used to represent the response for a typical participant. If a symbol

is above the line at RR=1, then that education group is higher than the geometric mean for a typical participant for that fish consumption or mercury intake variable. For example, women with greater than median education level of the participant's age (maroon open diamond of >Median of age) eat fish more frequently than typical women (horizontal line). The blue dot is the RR of the inverse of body weight. A RR less than one indicates higher body weight compared to a typical participant and a RR greater than one indicates lower body weight than typical.

There are significant differences (p<0.05) for all fish consumption and mercury intake variables by education except the mercury concentration in fish consumed (p=0.25). The percent of women who consumed fish in the previous 30 days and mercury intake per unit body weight increase with increasing known level of education. Women with less than median level education have the lowest mercury intake per body weight: they generally eat fish less frequently, eat a smaller meal size, and have the highest body weight. The proportion of women with unknown education level who consumed fish in the previous 30 days is the highest among all groups. They eat fish more frequently compared to women with less than median level education and median level education; however, their meal size and mercury concentration in fish consumed are the lowest, resulting in relative low mercury intake per unit body weight.

Figure 8.Relative ratios and 95 percent confidence intervals from models predicting fish
consumption and mercury intake variables by education, NHANES 2013–March 2020

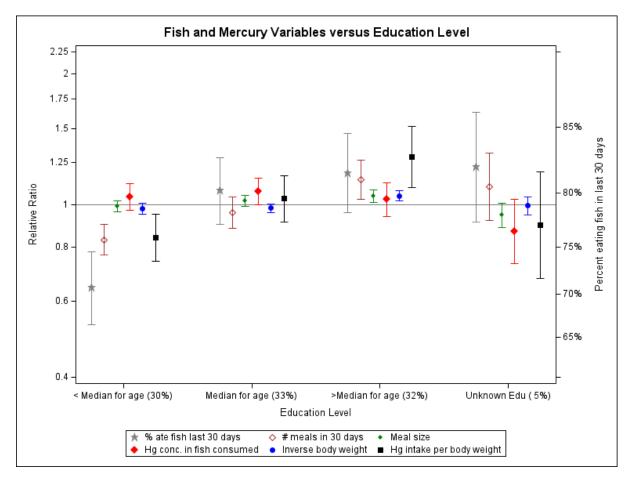
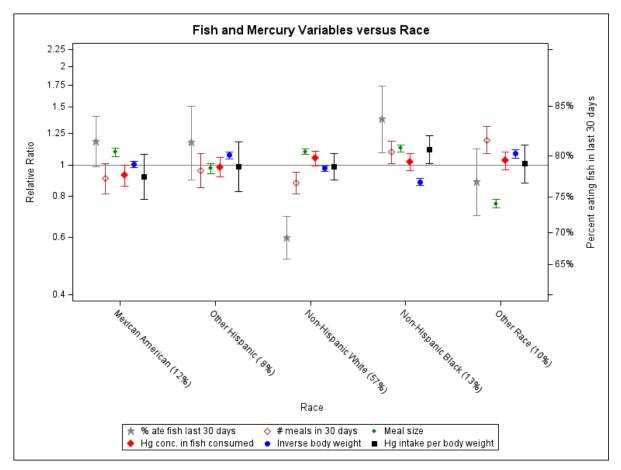
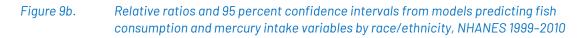
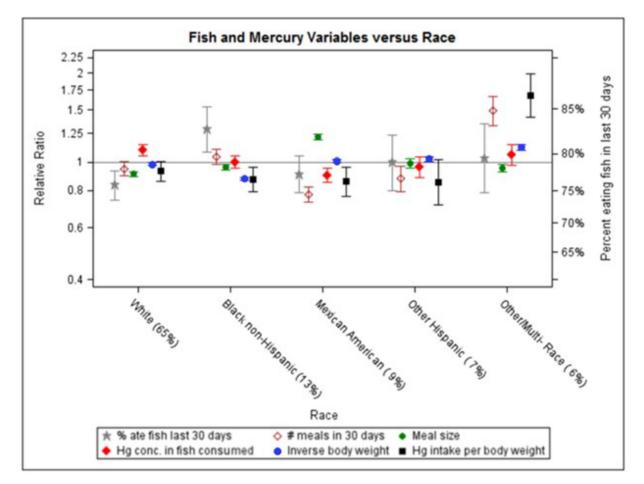


Figure 9a presents the RRs for fish consumption and mercury intake by race/ethnicity of NHANES 2013-March 2020. The p-values testing the overall significance of race/ethnicity in all models is p<0.0001, except for the model of mercury concentration of the fish consumed (p=0.21) and mercury intake per unit body weight (p=0.28). The percent of non-Hispanic Black women who consumed fish in the previous 30 days is higher compared to other racial/ethnic groups while consumption among non-Hispanic White women is lower, which is consistent with that found in NHANES 1999-2010 study (United States Environmental Protection Agency, 2013) presented in Figure 9b.









In general, non-Hispanic Black women consume the largest meal size, eat fish more frequently, consume fish with higher mercury concentration, and have higher body weights than a typical woman of childbearing age, resulting in the highest mercury intake per unit body weight of the racial/ethnic groups in NHANES 2013-March 2020. In NHANES 1999-2010 (Figure 9b), non-Hispanic Black women have similar distributions compared to a typical woman of childbearing age except they consume smaller than typical meal size, resulting in a less than typical mercury intake per unit body weight.

In NHANES 2013-March 2020, women who identify as "Other Race" eat fish the most frequently, consume fish with higher mercury concentration, have lower than typical body weight; however, while they consume the smallest meal size, they have a slightly higher than typical mercury intake per unit body weight. In NHANES 1999-2010, women of "Other Race" have the same trends in these fish consumption and mercury intake variables. With a close to typical meal size, they have the largest mercury intake per unit body weight of all racial/ethical groups.

In NHANES 2013-March 2020, non-Hispanic White women generally consume fish the least frequently and have higher than typical body weight; however, they consume larger meals, consume fish with higher than typical concentration of mercury, resulting in a close to typical mercury intake per unit body weight. In NHANES 1999-2010, the fish consumption and mercury intake variables of non-Hispanic White women follow the same trends except they consumed smaller than typical meal size and yield a less than typical mercury intake per unit body weight.

In NHANES 2013-March 2020, Mexican American women consume larger meal sizes; however, they eat less frequently and consumed fish with lower concentration of mercury, resulting in the lowest mercury intake per unit body weight. Other Hispanic women consume fish less frequently, eat smaller meal size, consume fish with lower contraction of mercury, and have lower body weight than a typical woman of childbearing age, resulting in a close to typical mercury intake per unit body weight. The trends of the fish consumption and mercury intake variables compared to values of typical women of childbearing age are similar in Mexican American and other-Hispanic participants of NHANES 1999-2010 (Figure 9b).

Figure 10 presents the RRs for fish consumption and mercury intake variables by income of NHANES 2013-March 2020 survey releases. There are significant differences (p=0.0004) for the proportion of women who consumed fish in the previous 30 days by income. Higher income is associated with higher proportion of women who consumed fish in the previous 30 days by income. Higher income fish the most frequently, consume larger meal sizes and fish with higher concentration of mercury, and have less than typical body weight, resulting in the highest mercury intake per unit body weight. In the 2013 study of trends in fish consumption among women of childbearing age using NHANES 1999-2010 data (United States Environmental Protection Agency, 2013), income is categorized with different survey variables and therefore not comparable to this study.

Figure 10.Relative ratios and 95 percent confidence intervals from models predicting fish
consumption and mercury intake variables by income, NHANES 2013–March 2020

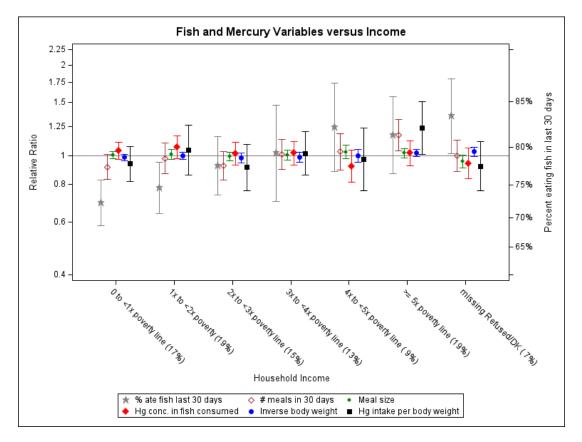
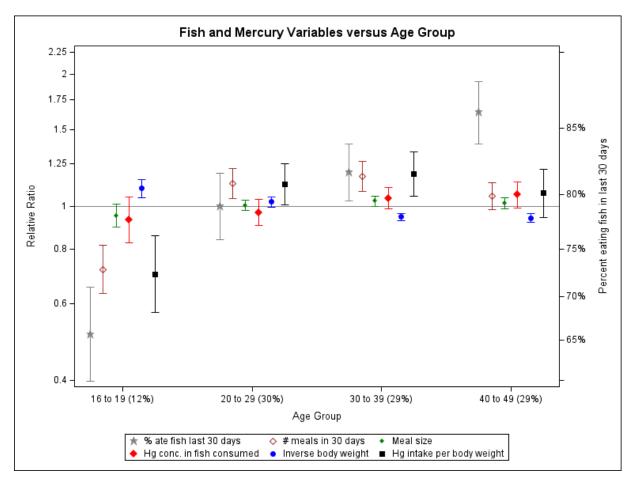
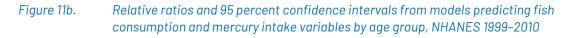
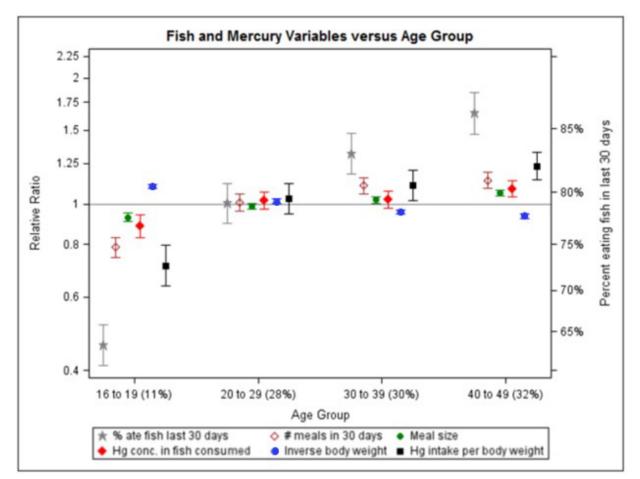


Figure 11a presents the RRs for fish consumption and mercury intake by age group. The pvalues testing the overall significance of age group in all models are less than 0.05, except for the models of mercury concentration of the fish consumed (p=0.16) and amount of meal consumed (p=0.29). The proportion of women who consumed fish in the previous 30 days increases with age. As observed in a previous study on trends in blood mercury concentrations and fish consumption among U.S. women of childbearing Age NHANES 1999-2010 (Birch *et al.*, 2014; United States Environmental Protection Agency, 2013) and presented in Figure 11b, women 16-19 years eat fish the least frequently, consume the smallest meal size, have the lowest body weight, and consume fish with the lowest concentration of mercury, resulting in the lowest mercury intake per unit body weight. The value of mercury intake per unit body weight increases with age to the highest level in women 30-39 years and then drops in women 40-49 years to a level comparable to women 20-39 years. Based on NHANES 1999-2010 study (United States Environmental Protection Agency, 2013) women of 40-49 years have the highest intakes of mercury per unit body weight.

Figure 11a.Relative ratios and 95 percent confidence intervals from models predicting fish
consumption and mercury intake variables by age group, NHANES 2013–March 2020







DISCUSSION AND CONCLUSIONS 4

Trends in geometric mean blood MeHg concentrations over time and by demographic characteristics

The results of this study indicate that there is a significant (p<0.0001, <u>Table A.3</u>) decreasing trend of blood MeHg concentrations over time across the ten NHANES survey releases (see Figure 1 and Table 3) after controlling for demographic characteristics. The geometric mean blood MeHg concentration is highest in NHANES survey release 1999-2000, then declines to the lowest in NHANES survey release 2017-March 2020. Additionally, the geometric mean blood MeHg concentration in the 2005-2006 release is slightly higher than that in the 2003-2004 and 2007-2008 releases, while the geometric mean blood MeHg in the 2009-2010 release is slightly higher than those in the 2007-2008 and 2011-2012 releases. This does not appear to be a meaningful increase and is likely due to the fluctuations in the data. The geometric mean blood MeHg concentration in NHANES 1999-2000 is 1.94 times higher than the geometric mean in NHANES 2017-March 2020 data, representing a 48 percent decrease between NHANES 1999-2000 and 2017-March 2020. Similar decreasing trends in blood MeHg concentration over time are found in NHANES 1999-2010 (Birch et al., 2014; United States Environmental Protection Agency, 2013). The studies using NHANES 1999-2010 found that the linear time trend in the mean of blood MeHg concentrations was statistically significant (p=0.006), but there was no significant trend from 2001 to 2010 (p=0.74). With additional NHANES survey releases in this study, excluding survey release 1999-2000 does not change the significance of the linear trend from 2001-March 2020.

The percentages of women of childbearing age with blood MeHg concentrations over 5.8 μ g/L in 1999-2000 is about 3.5 times that found in 2001-March 2020, representing a 71 percent decrease. There is a significant difference between the survey releases for blood MeHg concentrations (Rao-Scott Chi-square p<0.001), with 1999-2000 having approximately three times the amount of women with levels over 5.8 μ g/L compared to the other NHANES releases. However, no significance difference (p=0.12) is found between the survey releases after removing survey release NHANES 1999-2000. Similar patterns were found in previous studies of NHANES 1999-2010 (Birch *et al.*, 2014; United States Environmental Protection Agency, 2013).

There are significantly (p<0.05) decreasing trends of blood MeHg concentrations over time by most demographic characteristics across these NHANES survey releases 1999-2004, 2005-2010, 2011-2016, and 2017-March 2020. The geometric mean blood MeHg concentration in

women of "Other Race" in 1999-2004 is 1.82 times higher than the geometric mean for the same demographic in 2017-March 2020, representing a 45 percent decrease between NHANES 1999-2004 and 2017-March 2020. Within each of these survey release periods, higher blood MeHg concentrations are observed with increasing age, ratio of family income to poverty, education level, and among participants who reported their race as "Other Race."

Association of blood MeHg concentration with intake and demographic characteristics

A multivariable model is used to investigate the association between MeHg concentration and TUI through fish consumption and other demographic characteristics. The model found the following factors to be significantly associated with blood MeHg are (1) TUI through fish consumption (p<0.0001), (2) education (p=0.0007), (3) race/ethnicity (p<0.0001), (4) NHANES survey release (p<0.0001), (5) log-transformed hematocrit (p=0.005), and (6) log-transformed body weight (p=0.002). In addition, the rate of increase in blood MeHg concentration due to usual intake of mercury varies by education, race/ethnicity, and log-transformed body weight. Household income is marginally significant at the 5 percent level (p=0.054) in predicting blood MeHg concentrations.

Geographic differences in blood mercury

The geographic findings of this study (<u>Appendix B</u>) are similar to what was found in previous studies (Cusack *et al.*, 2017; Mahaffey *et al.*, 2009). There are geographic differences in blood MeHg concentrations with higher levels among residents of coastal counties compared to non-coastal counties. Residents of the Northeast region have the highest blood MeHg concentrations of the four regions and residents of the Midwest have the lowest blood MeHg concentrations.

Trends in frequency of fish consumption and the association of fish consumption and mercury intake with demographic characteristics

One limitation of this study is that the analysis of fish consumption over the previous 30 days used data from NHANES 2013-March 2020, and results were compared to a previous study using NHANES 1999-2010 (United States Environmental Protection Agency, 2013). NHANES 2011-2012 was not published during the previous study and was not included in the current study, leaving a gap between the two studies. This gap limits the ability to investigate the trend in fish consumption from 1999 to March 2020.

Blood MeHg concentrations are positively associated with the frequency of fish consumption in both NHANES 2013-March 2020 and NHANES 1999-2010. There are no significant differences in the reported frequency of consumption of fish over the previous 30 days across NHANES survey releases 2013-March 2020 while significant differences of reported frequency of consumption are found across the six NHANES survey releases between 1999 and 2010 (United States Environmental Protection Agency, 2013). The association between the frequency of fish consumption and demographic characteristics are consistent between NHANES 1999-2000 and NHANES 2013-March 2020, respectively. Women of older age and with higher income tend to eat fish more frequently. Individuals of "Other Race" eat fish more frequently compared to Mexican American, non-Hispanic White, non-Hispanic Black, and other Hispanic. Women with higher education levels are found to eat fish more frequently.

The estimated amounts of total fish eaten over the previous 30 days in NHANES 2013-March 2020 are at the higher end of those found in NHANES 1999-2010. The estimated mercury intake from total fish consumption over the previous 30 days in NHANES 2013-March 2020 are lower than those found in NHANES 1999-2010. This suggests that women of childbearing age are probably choosing to eat fish with lower mercury concentration leading to lower estimates of mercury intake per unit body weight in NHANES 2013-March 2020 compared to those found in NHANES 1999-2010.

In addition to continuing to monitor the time trend in blood MeHg and fish consumption, a future study on the geographic distribution of blood MeHg using NHANES releases of 2013-2014 and later would be a valuable supplement to this study, helping to investigate trends in geographic differences in blood MeHg.

QUALITY CONTROL/QUALITY ASSURANCE 5

This section details the steps that were taken to ensure the quality of the results.

The fish tissue mercury concentration database developed for previous EPA studies on trends in blood mercury concentrations and fish consumption was updated with data on fish samples collected from 2013 to 2020. Data were downloaded when they were available online. Other data extracted from peer reviewed journal articles and reports were checked by a second individual to ensure all information were correct.

The NHANES 2003-2010 24-hour recall data were processed to extract all reports of fish consumption for the 2014 U.S. EPA report (United States Environmental Protection Agency, 2014), *Estimated Fish Consumption Rates for the U.S. Population and Selected Subpopulations (NHANES 2003-2010)*. At that time, the processing was done independently by two individuals and results were compared. A final program with macros applicable to all NHANES releases in general was created to process NHANES 24-hour dietary recall data. In the study of the geographic distributions of blood mercury concentration (Appendix B) of NHANES 1999-2012, additions were made to that code to include processing of the NHANES 1999-2000, 2001-2002, and 2011-2012 data, based on the code for the NHANES 2003-2010 data. The current analysis is built upon these previous analyses with the addition of NHANES 2013-2014, 2015-2016, and 2017-March 2020 data.

This analysis utilized the software created for the estimation of usual fish consumption rates, the EPA Method, developed for <u>Estimated Fish Consumption Rates for the U.S. Population</u> <u>and Selected Subpopulations (NHANES 2003-2010)</u> (United States Environmental Protection Agency, 2014). This software has previously undergone quality checks. This analysis also utilized the INDIVINT macro, which is one of three NCI Method programs. It is available from NCI.

The imputation method was evaluated using the NHANES 2011-2012 data on methyl and ethyl mercury.

The models used a large number of potential predictors. These were very useful in avoiding misleading results that may be found in simpler models that do not account for these interrelationships.

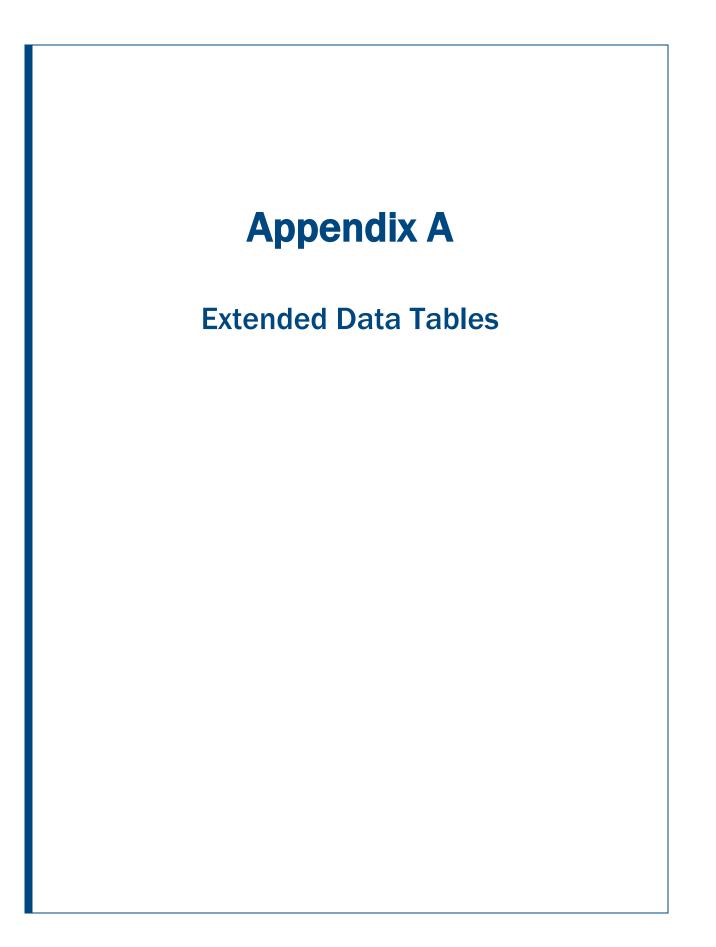


- Akinbami, L. J.; Chen, T. C.; Davy, O.; Ogden, C. L.; Fink, S.; Clark, J.; Riddles, M. K.; Mohadjer, L. K. (2022) <u>National Health and Nutrition Examination Survey, 2017–March 2020 Prepandemic File: Sample Design, Estimation, and Analytic Guidelines</u>. Vital Health Stat 1, 1–36. <u>https://dx.doi.org/10.15620/cdc:115434</u>.
- Birch, R. J.; Bigler, J.; Rogers, J. W.; Zhuang, Y.; Clickner, R. P. (2014) Trends in blood mercury concentrations and fish consumption among U.S. women of reproductive age, NHANES, 1999-2010. *Environ Res*, 133, 431-438. <u>https://doi.org/10.1016/j.envres.2014.02.001</u>.
- Björnberg, K. A.; Vahter, M.; Petersson-Grawé, K.; Glynn, A.; Cnattingius, S.; Darnerud, P. O.; Atuma, S.; Aune, M.; Becker, W.; Berglund, M. (2003) Methyl mercury and inorganic mercury in Swedish pregnant women and in cord blood: influence of fish consumption. *Environ Health Perspect*, 111, 637-641. <u>https://doi.org/10.1289/ehp.111-1241457</u>.
- Bramante, C. T.; Spiller, P.; Landa, M. (2018) Fish Consumption During Pregnancy: An Opportunity, Not a Risk. JAMA Pediatr, 172, 801-802. <u>https://doi.org/10.1001/jamapediatrics.2018.1619</u>.
- Caldwell KL, Mortensen ME, Jones RL, Caudill SP, Osterloh JD. Total blood mercury concentrations in the U.S. population: 1999-2006. Int J Hyg Environ Health. 2009 Nov;212(6):588-98. https://doi.org/10.1016/j.ijheh.2009.04.004.
- Center for Disease Control and Prevention. (2021a) Fourth national report on human exposure to environmental chemicals. Updated tables, March 2021: volume one, NHANES 1999-2010 in: National Center for Environmental Health. Division of Laboratory, S.; National Center for Health, S.; National, H.; Nutrition Examination, S. (Eds.), Atlanta, GA.
- Center for Disease Control and Prevention. (2021b) Fourth national report on human exposure to environmental chemicals. Updated tables, March 2021 : volume two, NHANES 2011-2016 in: National Center for Environmental Health. Division of Laboratory, S.; National Center for Health, S.; National, H.; Nutrition Examination, S. (Eds.). Center for Disease Control and Prevention, Atlanta, GA.
- Center for Disease Control and Prevention, N. C. f. H. S. N. (2013) About the National Health and Nutrition Examination Survey. U.S. Department of Health and Human Services, Centers for Disease Control and Prevention, Hyattsville, MD.
- Clarkson, T. W.; Magos, L. (2006) The toxicology of mercury and its chemical compounds. *Crit Rev Toxicol*, 36, 609-662. <u>https://doi.org/10.1080/10408440600845619</u>
- Cusack, L. K.; Smit, E.; Kile, M. L.; Harding, A. K. (2017) Regional and temporal trends in blood mercury concentrations and fish consumption in women of childbearing age in the United States using NHANES data from 1999-2010. *Environ Health*, 16, 10. <u>https://doi.org/10.1186/s12940-017-0218-4</u>.
- EPA-FDA. (2021) Advice about Eating Fish and Shellfish in: United States Environmental Protection Agency and United States Food and Drug Administration (Ed.). Available at: <u>https://www.epa.gov/choose-</u><u>fish-and-shellfish-wisely/epa-fda-advice-about-eating-fish-and-shellfish</u>
- Golding, J.; Gregory, S.; Iles-Caven, Y.; Hibbeln, J.; Emond, A.; Taylor, C. M. (2016) Associations between prenatal mercury exposure and early child development in the ALSPAC study. *Neurotoxicology*, 53, 215-222. <u>https://doi.org/10.1016%2Fj.neuro.2016.02.006</u>.
- Imanse, S. M.; Anchor, C. L.; Anchor, G. C.; Landolfi, J. A.; Kinsel, M. J.; Levengood, J. M.; Delaney, M. A.; Terio, K. A. (2022) Pathologic impacts of contaminants in freshwater fish of Cook County IL. Aquat Toxicol, 242, 106043. <u>https://doi.org/10.1016/j.aquatox.2021.106043</u>.

- Janssen, S. E.; Hoffman, J. C.; Lepak, R. F.; Krabbenhoft, D. P.; Walters, D.; Eagles-Smith, C. A.; Peterson, G.; Ogorek, J. M.; DeWild, J. F.; Cotter, A.; Pearson, M.; Tate, M. T.; Yeardley, R. B., Jr.; Mills, M. A. (2021) Examining historical mercury sources in the Saint Louis River estuary: How legacy contamination influences biological mercury levels in Great Lakes coastal regions. *Sci Total Environ*, 779, 146284. <u>https://doi.org/10.1016/j.scitotenv.2021.146284</u>.
- Janssen, S. E.; Riva-Murray, K.; DeWild, J. F.; Ogorek, J. M.; Tate, M. T.; Van Metre, P. C.; Krabbenhoft, D. P.; Coles, J. F. (2019) Chemical and Physical Controls on Mercury Source Signatures in Stream Fish from the Northeastern United States. *Environ Sci Technol*, 53, 10110-10119. <u>https://doi.org/10.1021/acs.est.9b03394</u>.
- Johnson, C. L.; Paulose-Ram, R.; Ogden, C. L.; Carroll, M. D.; Kruszon-Moran, D.; Dohrmann, S. M.; Curtin, L. R. (2013) National health and nutrition examination survey: analytic guidelines, 1999-2010. *Vital Health Stat 2*, 1-24. <u>https://www.cdc.gov/nchs/data/series/sr_02/sr02_161.pdf</u>
- Kipnis, V.; Midthune, D.; Buckman, D. W.; Dodd, K. W.; Guenther, P. M.; Krebs-Smith, S. M.; Subar, A. F.; Tooze, J. A.; Carroll, R. J.; Freedman, L. S. (2009) Modeling data with excess zeros and measurement error: application to evaluating relationships between episodically consumed foods and health outcomes. *Biometrics*, 65, 1003-1010. <u>https://doi.org/10.1111/j.1541-0420.2009.01223.x</u>.
- Lepak, R. F., Bartz, K.K., Ogorek, J.M., Tate, M.T., DeWild, J.F., and Janssen, S.E. (2022) Assessment of mercury sources in Alaskan lake food webs: U.S. Geological Survey data release, in: United States Geological Survey (Ed.).
- Louisiana Department of Environmental Quality. (2016-2020) Mercury Contaminant Levels in Louisiana Biota in: Quality, L. D. o. E. (Ed.).
- Mahaffey, K. R.; Clickner, R. P.; Jeffries, R. A. (2009) Adult women's blood mercury concentrations vary regionally in the United States: association with patterns of fish consumption (NHANES 1999-2004). Environ Health Perspect, 117, 47-53. <u>https://doi.org/10.1289%2Fehp.11674</u>.
- Mahaffey, K. R.; Clickner, R. P.; Bodurow, C. C. (2004) Blood organic mercury and dietary mercury intake: National Health and Nutrition Examination Survey, 1999 and 2000. *Environ Health Perspect*, 112, 562– 570. <u>https://doi.org/10.1289%2Fehp.6587</u>.
- Malinowski, C. R. (2019) High mercury concentrations in Atlantic Goliath Grouper: Spatial analysis of a vulnerable species. *Mar Pollut Bull*, 143, 81-91. <u>https://doi.org/10.1016/j.marpolbul.2019.04.006</u>.
- McGee, B. N. R., D.L.; Pribil, M. (2020) Hg Concentrations of Fish Tissue Samples in the Vicinity of Yellow Pine, Idaho in: United States Geological Survey (Ed.). Available at <u>https://www.usgs.gov/data/hg-</u> <u>concentrations-fish-tissue-samples-vicinity-yellow-pine-idaho</u>.
- Mebane, C. A., and Schmidt, C.G. (2019) Selenium and mercury in the Kootenai River, Montana and Idaho, 2018-2019: U.S. Geological Survey data release in: United States Geological Survey (Ed.). Available at <u>https://www.usgs.gov/data/selenium-and-mercury-kootenai-river-montana-and-idaho-2018-2019</u>.
- Melnyk, L. J.; Lin, J.; Kusnierz, D. H.; Pugh, K.; Durant, J. T.; Suarez-Soto, R. J.; Venkatapathy, R.; Sundaravadivelu, D.; Morris, A.; Lazorchak, J. M.; Perlman, G.; Stover, M. A. (2021) Risks from mercury in anadromous fish collected from Penobscot River, Maine. *Sci Total Environ*, 781, 146691. <u>https://doi.org/10.1016/i.scitotenv.2021.146691</u>.
- Mergler, D.; Anderson, H. A.; Chan, L. H.; Mahaffey, K. R.; Murray, M.; Sakamoto, M.; Stern, A. H.; Panel on Health, R.; Toxicological Effects of, M. (2007) Methylmercury exposure and health effects in humans: a worldwide concern. Ambio, 36, 3-11. <u>https://doi.org/10.1579/0044-7447(2007)36[3:meahei]2.0.co;2</u>.
- Murray, K. R.; Cleckner, L.B.; Razavi, N.R.; Richter, W. (2020) Fish mercury concentration data and ancillary data for streams and rivers across New York States (United States), 1969-2016, including environmental characteristics of selected locations sampled during 2007-16 in: United States Geological Survey (Ed.). Available at https://data.usgs.gov/datacatalog/data/USGS:5e1f563ce4b0ecf25c61ed64.

- National Research Council. (2000) *Toxicological Effects of Methylmercury*. The National Academies Press: Washington, DC. <u>https://doi.org/10.17226/9899</u>.
- New Hampshire Department of Environmental Services. (2018) Status and trends of mercury in fish tissue in New Hampshire waterbodies, 1992-2016. https://www.des.nh.gov/sites/g/files/ehbemt341/files/documents/2020-01/r-wd-17-22.pdf
- Park, S.; Lee, B. K. (2013) Strong positive associations between seafood, vegetables, and alcohol with blood and urinary arsenic levels in the Korean adult population. Arch Environ Contam Toxicol, 64 (1), 167-170. <u>https://doi.org/10.1007/s00244-012-9808-x</u>
- Rice, G.; Swartout, J.; Mahaffey, K.; Schoeny, R. (2000) Derivation of U.S. EPA's oral Reference Dose (RfD) for methylmercury. *Drug Chem Toxicol*, 23, 41-54. <u>https://doi.org/10.1081/dct-100100101</u>.
- Rothenberg, S. E.; Korrick, S. A.; Fayad, R. (2015) The influence of obesity on blood mercury levels for U.S. non-pregnant adults and children: NHANES 2007-2010. *Environ Res*, 138, 173-180. <u>https://doi.org/10.1016/j.envres.2015.01.018</u>.
- Sackett, D. K.; Drazen, J. C.; Popp, B. N.; Choy, C. A.; Blum, J. D.; Johnson, M. W. (2017) Carbon, Nitrogen, and Mercury Isotope Evidence for the Biogeochemical History of Mercury in Hawaiian Marine Bottomfish. Environ Sci Technol, 51, 13976-13984. <u>https://doi.org/10.1021/acs.est.7b04893</u>.
- Sanzo, J. M.; Dorronsoro, M.; Amiano, P.; Amurrio, A.; Aguinagalde, F. X.; Azpiri, M. A.; Spain, E. G. o. (2001) Estimation and validation of mercury intake associated with fish consumption in an EPIC cohort of Spain. *Public Health Nutr*, 4, 981-988.
- SAS Institute. (2016) Statistical Analysis System, 9.4 ed. https://support.sas.com/software/94/.
- Stratakis, N.; Conti, D. V.; Borras, E.; Sabido, E.; Roumeliotaki, T.; Papadopoulou, E.; Agier, L.; Basagana, X.;
 Bustamante, M.; Casas, M.; Farzan, S. F.; Fossati, S.; Gonzalez, J. R.; Grazuleviciene, R.; Heude, B.;
 Maitre, L.; McEachan, R. R. C.; Theologidis, I.; Urquiza, J.; Vafeiadi, M.; West, J.; Wright, J.; McConnell,
 R.; Brantsaeter, A. L.; Meltzer, H. M.; Vrijheid, M.; Chatzi, L. (2020) Association of Fish Consumption
 and Mercury Exposure During Pregnancy With Metabolic Health and Inflammatory Biomarkers in
 Children. JAMA Netw Open, 3, e201007. https://doi.org/10.1001%2Fjamanetworkopen.2020.1007.
- Svensson, B. G.; Schütz, A.; Nilsson, A.; Akesson, I.; Akesson, B.; Skerfving, S. (1992) Fish as a source of exposure to mercury and selenium. *Sci Total Environ*, 126, 61-74. <u>https://doi.org/10.1016/0048-9697(92)90484-a</u>.
- Taylor, D. L.; Calabrese, N. M. (2018) Mercury content of blue crabs (Callinectes sapidus) from southern New England coastal habitats: Contamination in an emergent fishery and risks to human consumers. *Mar Pollut Bull*, 126, 166-178. <u>https://doi.org/10.1016/j.marpolbul.2017.10.089</u>.
- Taylor, C.; Golding, J.; Emond, A. M. (2016) Blood mercury levels and fish consumption in pregnancy: Risks and benefits for birth outcomes in a prospective observational birth cohort. *International Journal of Hygiene and Environmental Health*, 219, 513-520. <u>https://doi.org/10.1016%2Fj.ijheh.2016.05.004</u>.
- Tooze, J. A.; Kipnis, V.; Buckman, D. W.; Carroll, R. J.; Freedman, L. S.; Guenther, P. M.; Krebs-Smith, S. M.; Subar, A. F.; Dodd, K. W. (2010) A mixed-effects model approach for estimating the distribution of usual intake of nutrients: the NCI method. *Stat Med*, 29, 2857-2868. <u>https://doi.org/10.1002/sim.4063</u>.
- Tooze, J. A.; Midthune, D.; Dodd, K. W.; Freedman, L. S.; Krebs-Smith, S. M.; Subar, A. F.; Guenther, P. M.; Carroll, R. J.; Kipnis, V. (2006) A new statistical method for estimating the usual intake of episodically consumed foods with application to their distribution. J Am Diet Assoc, 106, 1575-1587. <u>https://doi.org/10.1016/i.jada.2006.07.003</u>.
- United States Environmental Protection Agency. (2023) 2013-2014 National Rivers and Streams Assessment Fish Tissue Study in: United States Environmental Protection Agency (Ed.). Available at <u>https://www.epa.gov/choose-fish-and-shellfish-wisely/2013-2014-national-rivers-and-streams-assessment-fish-tissue-study</u>.

- United States Environmental Protection Agency. (2015) 2015 Great Lakes Human Health Fish Fillet Tissue Study in: United States Environmental Protection Agency (Ed.). Available at <u>https://www.epa.gov/choose-fish-and-shellfish-wisely/2015-great-lakes-human-health-fish-fillet-tissue-study</u>.
- United States Environmental Protection Agency. (2014) Estimated Fish Consumption Rates for the U.S. Population and Selected Subpopulations (NHANES 2003-2010). Report No. EPA-820-R-14-002; United States Environmental Protection Agency, Office of Science and Technology, Office of Water. Available at <u>https://www.epa.gov/sites/default/files/2015-01/documents/fish-consumption-rates-2014.pdf</u>.
- United States Environmental Protection Agency. (2013) *Trends in Blood Mercury Concentrations and Fish Consumption Among U.S. Women of Childbearing Age, NHANES 1999-2010.* Report No. EPA-823-R-13-002; U.S. Environmental Protection Agency: Washington D.C. Available at <u>https://www.epa.gov/sites/default/files/2018-11/documents/trends-blood-mercury-concentrations-</u> <u>report.pdf</u>.
- United States Environmental Protection Agency. (2001) Integrated Risk Information System Chemical Assessment Summary: Methylmercury (MeHg) CASRN 22967-92-6. United States Environmental Protection Agency, National Center for Environmental Assessment. Available at https://iris.epa.gov/ChemicalLanding/&substance_nmbr=73.
- United States Environmental Protection Agency. (1997) <u>Mercury Study Report to Congress, Volume IV: An</u> <u>Assessment of Exposure to Mercury in the United States</u>, Report No. EPA/452-R-97-006; USEPA Office of Air Quality Planning & Standards and Office of Research and Development. Washington, D.C.
- Washington State Department of Ecology. (2011-2016) Measuring Mercury Trends in Freshwater Fish in Washington State in: Ecology, W. S. D. o. (Ed.).
- Whitney, M. C. (2021) Metal accumulation in Lake Michigan prey fish: Influence of ontogeny, trophic position, and habitat. *Journal of Great Lakes research*, v. 47, pp. 1746-1755-2021 v.1747 no.1746. <u>https://doi.org/10.1016/j.jglr.2021.08.019</u>.
- Willacker, J. J., Eagles-Smith, C.A., Blazer, V.S. (2020) Total Mercury Concentrations in Smallmouth Bass from Chesapeake Bay Tributaries, USA Dataset, 2013-2017 in: United States Geological Survey (Ed.). Available at <u>https://data.usgs.gov/datacatalog/data/USGS:5d7ff0f3e4b0c4f70d04be43</u>.
- Wolff, S.; Brown, G.; Chen, J.; Meals, K.; Thornton, C.; Brewer, S.; Cizdziel, J. V.; Willett, K. L. (2016) Mercury concentrations in fish from three major lakes in north Mississippi: Spatial and temporal differences and human health risk assessment. *J Toxicol Environ Health A*, 79, 894-904. <u>https://doi.org/10.1080/15287394.2016.1194792</u>.



| Species group | µg Hg/g fish | Species group | µg Hg/g fish | Species group | µg Hg/g fish |
|-----------------------------------|-----------------|------------------------------|-----------------|--|-----------------|
| Swordfish | 0.739 | Snapper | 0.106 | Herring/Shad | 0.042 |
| Barracuda | 0.483 | Catfish | 0.090 | Smelt | 0.036 |
| Shark | 0.410 | Fish, not specified | 0.083 | Salmon | 0.036 |
| Ray | 0.410 | Haddock | 0.066 | Mullet | 0.032 |
| Mackerel | 0.311 | Cod | 0.065 | ³ Shellfish, not specified | 0.029 |
| Tuna (Fresh/Frozen) | 0.281 | Crab | 0.064 | Octopus | 0.017 |
| Pike | 0.229 | Croaker | 0.058 | Abalone | 0.016 |
| Pompano/Mahi Mahi | 0.190 | Trout | 0.056 | Sardine | 0.016 |
| Perch/Bass/Walleye/ Bluegill | 0.179 | Whitefish | 0.054 | Anchovy | 0.016 |
| Lobster | 0.173 | Crayfish | 0.053 | Mussel | 0.015 |
| Halibut | 0.170 | Whiting | 0.052 | Scallop | 0.015 |
| Rockfish/Redfish/Orange Roughy | 0.169 | Flatfish | 0.050 | Shrimp | 0.014 |
| Sea Bass | 0.165 | Squid | 0.049 | Oyster | 0.014 |
| Eel | 0.152 | Breaded Fish Products | 0.047 | Clam | 0.011 |
| ¹ Tuna | 0.123 | Conch | 0.047 | Pollock | 0.011 |
| Sturgeon | 0.122 | Snail | 0.047 | Jellyfish | 0.010 |
| Tuna (Canned) | 0.113 | ² Other shellfish | 0.047 | Tilapia | 0.010 |
| Carp/Sucker | 0.109 | Porgy/Scup | 0.046 | Caviar/Roe | 0.000 |

Table A.1. Fish tissue mercury concentrations used in analysis, by species

¹ Tuna, weighted mean of Tuna (Fresh/Frozen) and Tuna (Canned).

² Other shellfish, weighted mean of abalone, conch, jellyfish, octopus, squid, and snail.

³ Shellfish, NS (non-specified), weighted mean of abalone, clam, conch, crayfish, jellyfish, lobster, mussel, octopus, oyster, scallop, shrimp, squid, crab, and snail.

| | | Arithmetic | Geometric | | Sele | cted percentiles (9 | 5% CI) | |
|--|------|-------------------|-------------------|-------------------|-------------------|---------------------|-------------------|--------------------|
| | N | mean (95% CI) | mean (95% CI) | 25th | 50th | 75th | 90th | 95th |
| NHANES Survey Release | e | | | | | | | |
| 1999-2000 | 1632 | 1.83 (1.43, 2.22) | 0.93 (0.78, 1.12) | 0.41(0.31, 0.51) | 0.89(0.68, 1.09) | 1.88(1.16, 2.60) | 4.56 (3.25, 5.86) | 6.82 (3.57, 10.07) |
| 2001-2002 | 1799 | 1.30 (1.13, 1.47) | 0.72 (0.65, 0.79) | 0.34 (0.29, 0.38) | 0.70 (0.60, 0.79) | 1.46 (1.23, 1.70) | 2.92 (2.56, 3.28) | 4.42 (3.44, 5.39) |
| 2003-2004 | 1615 | 1.06 (0.91, 1.22) | 0.53 (0.45, 0.61) | 0.24 (0.19, 0.29) | 0.52 (0.40, 0.64) | 1.21(0.98, 1.44) | 2.60 (2.00, 3.19) | 3.77 (2.87, 4.67) |
| 2005-2006 | 1788 | 1.13 (0.99, 1.27) | 0.58 (0.51, 0.66) | 0.25 (0.18, 0.32) | 0.63 (0.52, 0.73) | 1.33 (1.07, 1.59) | 2.71(2.19, 3.23) | 3.99(2.93, 5.05 |
| 2007-2008 | 1486 | 0.99(0.84,1.13) | 0.52 (0.46, 0.59) | 0.25 (0.20, 0.29) | 0.51(0.43, 0.59) | 1.06 (0.82, 1.30) | 2.39 (1.81, 2.97) | 3.46 (2.70, 4.23 |
| 2009-2010 | 1780 | 1.18 (1.08, 1.29) | 0.67 (0.61, 0.73) | 0.31(0.26, 0.37) | 0.64 (0.53, 0.74) | 1.38 (1.15, 1.61) | 2.73 (2.41, 3.04) | 4.03 (3.50, 4.55 |
| 2011-2012 | 1428 | 0.93 (0.75, 1.11) | 0.52 (0.44, 0.60) | 0.23 (0.20, 0.26) | 0.45 (0.37, 0.54) | 0.98 (0.65, 1.32) | 2.32 (1.66, 2.99) | 3.31(2.22, 4.41 |
| 2013-2014 | 814 | 0.98(0.84, 1.12) | 0.50 (0.45, 0.55) | 0.23 (0.18, 0.28) | 0.43 (0.37, 0.49) | 0.94 (0.78, 1.11) | 2.21(1.62, 2.80) | 3.66 (3.13, 4.20 |
| 2015-2016 | 740 | 0.90 (0.80, 1.01) | 0.50 (0.45, 0.56) | 0.24 (0.21, 0.27) | 0.44 (0.37, 0.51) | 0.94 (0.77, 1.10) | 2.19 (1.71, 2.68) | 3.38 (2.79, 3.96 |
| 2017-2020 | 2154 | 0.87 (0.75, 0.98) | 0.48(0.43, 0.54) | 0.18 (0.15, 0.21) | 0.43 (0.36, 0.50) | 1.01(0.79, 1.22) | 1.99 (1.54, 2.44) | 3.06 (2.29, 3.83 |
| Race/Ethnicity | | | | | | | | |
| Mexican American | 3396 | 0.69 (0.65, 0.73) | 0.44 (0.42, 0.46) | 0.24 (0.23, 0.26) | 0.42(0.39, 0.45) | 0.80 (0.74, 0.85) | 1.44 (1.33, 1.56) | 2.09 (1.77, 2.41 |
| Other Hispanic | 1259 | 1.16 (1.00, 1.31) | 0.66 (0.62, 0.71) | 0.31(0.28, 0.34) | 0.65 (0.59, 0.71) | 1.39 (1.24, 1.54) | 2.60 (2.27, 2.92) | 3.68 (3.02, 4.34 |
| Non-Hispanic White | 5725 | 1.05 (0.98, 1.11) | 0.54 (0.51, 0.57) | 0.23 (0.22, 0.25) | 0.51(0.47, 0.54) | 1.17 (1.09, 1.24) | 2.54 (2.27, 2.82) | 3.84 (3.52, 4.17 |
| Non-Hispanic Black | 3497 | 1.13 (1.05, 1.21) | 0.66 (0.62, 0.70) | 0.32 (0.30, 0.34) | 0.64 (0.60, 0.69) | 1.27 (1.16, 1.37) | 2.38 (2.10, 2.66) | 3.61(3.19, 4.03 |
| Other Race | 1359 | 2.01(1.85, 2.17) | 0.95 (0.88, 1.03) | 0.34 (0.29, 0.38) | 0.94 (0.82, 1.06) | 2.50 (2.18, 2.83) | 5.27(4.64, 5.90) | 7.64 (6.26, 9.01 |
| Age, Years | | | | | | | | |
| 16 to 19 | 3349 | 0.63 (0.59, 0.68) | 0.36 (0.35, 0.38) | 0.16 (0.15, 0.17) | 0.33 (0.30, 0.35) | 0.69(0.63, 0.75) | 1.46 (1.32, 1.60) | 2.10 (1.94, 2.26) |
| 20 to 29 | 4111 | 0.97 (0.91, 1.03) | 0.52 (0.49, 0.55) | 0.23 (0.20, 0.25) | 0.49(0.45,0.53) | 1.12 (1.01, 1.22) | 2.34 (2.14, 2.55) | 3.43 (3.01, 3.84 |
| 30 to 39 | 3943 | 1.22 (1.13, 1.32) | 0.63 (0.60, 0.66) | 0.28 (0.26, 0.29) | 0.60 (0.56, 0.65) | 1.34 (1.22, 1.45) | 2.92 (2.54, 3.29) | 4.30 (3.79, 4.81 |
| 40 to 49 | 3833 | 1.27 (1.19, 1.34) | 0.68 (0.65, 0.72) | 0.32 (0.30, 0.34) | 0.65 (0.60, 0.69) | 1.39 (1.28, 1.51) | 3.03 (2.75, 3.30) | 4.31(4.00, 4.62 |
| Annual Income | | | | | | | | |
| 0 to <1x poverty line | 3814 | 0.79 (0.73, 0.85) | 0.44 (0.42, 0.47) | 0.22(0.20, 0.24) | 0.41(0.38, 0.45) | 0.84 (0.78, 0.90) | 1.67 (1.49, 1.84) | 2.74 (2.41, 3.07 |
| 1x to <2x poverty | 3556 | 0.83 (0.78, 0.88) | 0.46(0.44, 0.48) | 0.22(0.20, 0.24) | 0.43 (0.40, 0.46) | 0.90 (0.83, 0.97) | 1.76 (1.59, 1.94) | 2.74 (2.40, 3.07 |
| 2x to <3x poverty line | 1995 | 1.00 (0.92, 1.08) | 0.53 (0.50, 0.57) | 0.25 (0.23, 0.27) | 0.50 (0.46, 0.54) | 1.11(0.98, 1.24) | 2.40(2.17, 2.63) | 3.31(2.78, 3.85 |
| 3x to <4x poverty line | 1580 | 1.06 (0.98, 1.14) | 0.57 (0.53, 0.61) | 0.25 (0.21, 0.28) | 0.56 (0.50, 0.62) | 1.17 (1.05, 1.28) | 2.54 (2.18, 2.90) | 3.74 (3.33, 4.15 |
| 4x to <5x poverty line | 1102 | 1.15 (1.04, 1.27) | 0.62 (0.57, 0.66) | 0.29 (0.26, 0.32) | 0.59 (0.51, 0.68) | 1.25 (1.11, 1.38) | 2.71(2.19, 3.23) | 4.36 (3.35, 5.38 |
| >= 5x poverty line | 2051 | 1.66 (1.55, 1.78) | 0.89 (0.83, 0.94) | 0.39(0.34, 0.43) | 0.92 (0.84, 1.01) | 2.02 (1.80, 2.24) | 3.98 (3.66, 4.30) | 5.90 (5.03, 6.77 |
| Missing/Refused/ DK | 1138 | 1.25 (1.08, 1.42) | 0.64 (0.58, 0.71) | 0.27 (0.24, 0.31) | 0.57 (0.50, 0.65) | 1.47 (1.14, 1.79) | 3.01(2.20, 3.83) | 4.51(3.28, 5.74 |
| Education | | | | | | | | |
| <median education<br="">for age</median> | 6139 | 0.89(0.83, 0.94) | 0.47 (0.45, 0.49) | 0.22 (0.21, 0.24) | 0.43 (0.40, 0.46) | 0.92 (0.86, 0.98) | 1.83 (1.67, 1.99) | 2.91(2.58, 3.24 |
| Median education for age | 5313 | 0.98 (0.93, 1.03) | 0.55(0.52, 0.57) | 0.25 (0.24, 0.27) | 0.53 (0.49, 0.56) | 1.13 (1.05, 1.21) | 2.25 (2.04, 2.46) | 3.25 (2.98, 3.53 |
| >Median education for age | 3411 | 1.52 (1.42, 1.62) | 0.80 (0.75, 0.85) | 0.35 (0.31, 0.38) | 0.79 (0.72, 0.87) | 1.82 (1.63, 2.01) | 3.78 (3.53, 4.04) | 5.33 (4.69, 5.97 |
| Unknown education | 373 | 0.51(0.42, 0.61) | 0.33 (0.29, 0.38) | 0.16 (0.15, 0.17) | 0.27 (0.21, 0.34) | 0.58 (0.45, 0.72) | 1.10 (0.65, 1.55) | 1.61(,) |

Table A.2. Distribution of blood MeHg concentrations (µg/L), by NHANES survey releases, age, income, and race/ethnicity, women aged 16-49 years, NHANES 2013-March 2020

^a Missing confidence interval because of stratum with single sampling unit.

| MeHg concentration | ons | | | | | | |
|--|----------|---------|---------|----------|--------|--------|----------|
| Dependent variable | | | Log-tra | nsformed | MeHg | | |
| Parameter | Estimate | LCL | UCL | tValue | Probt | fValue | ProbF |
| Intercept | 0.8410 | 0.5293 | 1.1528 | 5.3297 | 0.0000 | | |
| Factors affecting the | | | | | | | |
| intercept | | | | | | | |
| Education | | | | | | 5.6554 | 0.0007 |
| < Median for age | -0.2932 | -0.5570 | -0.0293 | -2.1952 | 0.0297 | | |
| >Median for age | 0.2599 | -0.0359 | 0.5558 | 1.7356 | 0.0846 | | |
| Unknown education | 0.0788 | -0.5782 | 0.7357 | 0.2369 | 0.8131 | | |
| Median for age | -0.0455 | -0.2813 | 0.1903 | -0.3814 | 0.7034 | | |
| Annual Income | | | | | | 2.0628 | 0.0541 |
| 0 to <1x poverty line | -0.0968 | -0.1848 | -0.0088 | -2.1742 | 0.0312 | | |
| 1x to <2x poverty | -0.1135 | -0.1999 | -0.0270 | -2.5935 | 0.0104 | | |
| 2x to <3x poverty line | 0.0021 | -0.1169 | 0.1212 | 0.0353 | 0.9719 | | |
| 3x to <4x poverty line | -0.0285 | -0.1468 | 0.0898 | -0.4758 | 0.6349 | | |
| 4x to <5x poverty line | 0.1043 | -0.0336 | 0.2422 | 1.4936 | 0.1373 | | |
| >= 5x poverty line | 0.1345 | 0.0076 | 0.2614 | 2.0942 | 0.0379 | | |
| Missing/Refused/DK | -0.0022 | -0.1654 | 0.1610 | -0.0262 | 0.9791 | | |
| Race/Ethnicity | | | | | | 6.6549 | 2.37E-05 |
| Mexican American | -0.4295 | -0.6110 | -0.2480 | -4.6743 | 0.0000 | | |
| Other Hispanic | 0.1070 | -0.1179 | 0.3318 | 0.9399 | 0.3488 | | |
| Non-Hispanic White | 0.4437 | 0.1775 | 0.7099 | 3.2926 | 0.0012 | | |
| Non-Hispanic Black | -0.0661 | -0.2674 | 0.1352 | -0.6487 | 0.5175 | | |
| Other Race | -0.0551 | -0.3539 | 0.2437 | -0.3642 | 0.7162 | | |
| NHANES Survey release | | | | | | 6.0979 | 1.28E-08 |
| 1999-2000 | 0.4773 | 0.3125 | 0.6422 | 5.7200 | 0.0000 | | |
| 2001-2002 | 0.0918 | -0.0898 | 0.2734 | 0.9991 | 0.3193 | | |
| 2003-2004 | -0.1155 | -0.2421 | 0.0111 | -1.8028 | 0.0734 | | |
| 2005-2006 | -0.0901 | -0.2296 | 0.0495 | -1.2754 | 0.2041 | | |
| 2007-2008 | 0.0132 | -0.1029 | 0.1292 | 0.2243 | 0.8228 | | |
| 2009-2010 | 0.1350 | 0.0166 | 0.2534 | 2.2534 | 0.0257 | | |
| 2011-2012 | -0.1197 | -0.2953 | 0.0559 | -1.3462 | 0.1802 | | |
| 2013-2014 | -0.1850 | -0.2957 | -0.0742 | -3.3010 | 0.0012 | | |
| 2015-2016 | -0.0640 | -0.2403 | 0.1123 | -0.7169 | 0.4745 | | |
| 2017-2020 | -0.1431 | -0.2852 | -0.0011 | -1.9909 | 0.0483 | | |
| Age, decade (centered) | 0.0212 | -0.0290 | 0.0714 | 0.8328 | 0.4063 | | |
| Log-transformed body | 0 7/.70 | -1.2101 | -0.2843 | -3.1887 | 0.0017 | | |
| weight (centered) | -0.7472 | -1.2101 | -0.2043 | -3.100/ | 0.0017 | | |
| Log-transformed hematocrit (centered) | 1.8543 | 0.5673 | 3.1413 | 2.8462 | 0.0050 | | |
| Age, decade (centered)*log- transformed body weight | -0.0769 | -0.2327 | 0.0789 | -0.9753 | 0.3310 | | |
| (centered) | | | | | | | |

Table A.3.Regression parameter estimates and p-values from models predicting log-transformed blood
MeHg concentrations

| Dependent variable | Log-transformed MeHg Estimate LCL UCL tValue Probt fValue P | | | | | | | | | |
|---|--|------------|----------|---------|--------|---------|----------|--|--|--|
| Parameter | Estimate | LCL | UCL | tValue | Probt | fValue | ProbF | | | |
| Factors Predicting the Slope for | Transformed | Usual Fish | Intake | | | | | | | |
| Transformed usual fish | 1.5657 | 1.3056 | 1.8258 | 11.8926 | 0.0000 | | | | | |
| intake | 1.0007 | 1.0000 | 1.0200 | 11.0020 | 0.0000 | | | | | |
| Transformed usual fish | 0.0/00 | 1 00 / / | 0.0001 | 7 7070 | 0 0000 | | | | | |
| intake*log-transformed body weight (centered) | -0.6468 | -1.0244 | -0.2691 | -3.3839 | 0.0009 | | | | | |
| Transformed usual fish | | | | | | | | | | |
| intake*log-transformed | 0.8567 | -0.2088 | 1.9222 | 1.5885 | 0.1143 | | | | | |
| hematocrit (centered) | 0.0007 | 0.2000 | 1.0222 | 1.0000 | 0.1110 | | | | | |
| Transformed usual fish | | | | | | 0 (070 | 0.075.07 | | | |
| intake*Race/Ethnicity | | | | | | 8.4232 | 8.63E-07 | | | |
| Transformed usual fish | -0.2898 | -0.4434 | -0.1361 | -3.7279 | 0.0003 | | | | | |
| intake*Mexican American | 0.2000 | 0.1101 | 0.1001 | 0.7270 | 0.0000 | | | | | |
| Transformed usual fish | -0.0686 | -0.2514 | 0.1141 | -0.7422 | 0.4591 | | | | | |
| intake*Other Hispanic Transformed usual fish | | | | | | | | | | |
| intake*Non-Hispanic White | 0.5229 | 0.3130 | 0.7328 | 4.9212 | 0.0000 | | | | | |
| Transformed usual fish | | | | | | | | | | |
| intake*Non-Hispanic Black | -0.1758 | -0.3507 | -0.0010 | -1.9865 | 0.0488 | | | | | |
| Transformed usual fish | 0.0113 | -0.2342 | 0.2569 | 0.0911 | 0.9276 | | | | | |
| intake*Other Race | 0.0113 | -0.2342 | 0.2009 | 0.0911 | 0.9270 | | | | | |
| Transformed usual fish | | | | | | 3.7225 | 0.0109 | | | |
| intake*Education | | | | | | | | | | |
| Transformed usual fish | -0.1855 | -0.3953 | 0.0243 | -1.7470 | 0.0827 | | | | | |
| intake*< Median for age Transformed usual fish | | | | | | | | | | |
| intake*>Median for age | 0.1903 | -0.0485 | 0.4290 | 1.5741 | 0.1175 | | | | | |
| Transformed usual fish | 0.0001 | 0 500 / | 0 55 (5 | | | | | | | |
| intake*Unknown Edu | 0.0261 | -0.5024 | 0.5547 | 0.0977 | 0.9223 | | | | | |
| Transformed usual fish | -0.0309 | -0.2208 | 0.1589 | -0.3221 | 0.7479 | | | | | |
| intake*Median for age | | | | | | | | | | |
| Τυιν | -0.0177 | -0.1347 | 0.0993 | -0.2986 | 0.7657 | | | | | |

Note: p-values in bold are for the F-Test for differences across categories in categorical predictors.

| D | | | | - | Percent (standard | derror) | | |
|----------------|------------------------|-------|------------|------------|-------------------|------------|------------|------------|
| Parameter | | N | 0 times | 1 time | 2 times | 3 times | 4-5 times | 6+ times |
| NHANES Survey | Release | | | | | | | |
| | 2013-2014 | 814 | 27.2(2.4) | 16.0(2.0) | 11.6 (1.2) | 8.9(1.1) | 9.8(1.4) | 26.4(2.2) |
| Total Fish | 2015-2016 | 740 | 26.5 (2.1) | 15.2(1.3) | 10.1(2.0) | 7.8(1.1) | 14.9(1.4) | 25.5 (2.5) |
| | 2017-March 2020 | 2,154 | 25.5(1.6) | 13.8(1.1) | 12.4 (1.2) | 9.0(0.9) | 13.3 (1.1) | 26.1(2.1) |
| | 2013-2014 | 814 | 39.3(2.6) | 19.0 (2.2) | 11.2 (1.2) | 7.4 (1.1) | 8.3(1.4) | 14.7(1.4) |
| Finfish Only | 2015-2016 | 740 | 38.0(1.9) | 17.1(1.7) | 12.6(1.6) | 8.7(1.4) | 10.9(1.7) | 12.8(1.5) |
| | 2017-March 2020 | 2,154 | 37.8(1.3) | 15.4 (1.2) | 14.6(1.0) | 8.7(0.6) | 10.2 (1.1) | 13.3(0.9) |
| | 2013-2014 | 814 | 49.3(2.6) | 16.0(1.2) | 13.3 (1.0) | 6.1(1.1) | 6.9(1.3) | 8.4(1.6) |
| Shellfish Only | 2015-2016 | 740 | 49.1(4.0) | 17.9 (2.2) | 12.3 (1.4) | 5.6(1.2) | 6.2(0.9) | 9.0(1.3) |
| | 2017-March 2020 | 2,154 | 47.6 (3.1) | 17.7(1.1) | 11.2 (0.9) | 8.2(1.2) | 8.1(1.4) | 7.3 (1.0) |
| Income | | | | | | | | |
| | 0 to <1x poverty line | 876 | 33.5(1.8) | 16.1(1.4) | 11.3 (1.4) | 11.1(1.5) | 11.1(1.4) | 16.9(1.9) |
| | 1x to <2x poverty line | 856 | 30.8(2.1) | 15.9(1.9) | 9.6(1.4) | 9.5 (1.3) | 11.2 (1.3) | 23.1(2.0) |
| | 2x to <3x poverty line | 516 | 27.0(2.6) | 17.3 (2.3) | 11.3 (1.8) | 11.2 (1.8) | 10.9(1.8) | 22.3(2.1) |
| Total Fish | 3x to <4x poverty line | 368 | 25.2(3.9) | 14.0(2.4) | 12.9(2.3) | 6.0(1.3) | 15.2(2.2) | 26.8(4.1) |
| | 4x to <5x poverty line | 251 | 20.0 (3.1) | 13.1(2.5) | 15.3 (3.3) | 8.8(2.3) | 10.5(2.3) | 32.2(4.4) |
| | >= 5x poverty line | 498 | 21.0(2.5) | 11.7 (2.1) | 10.1(1.6) | 5.8(1.4) | 15.1(1.9) | 36.4 (2.9) |
| | Missing/Refused/DK | 343 | 19.5 (2.5) | 15.9 (2.5) | 13.9(2.9) | 7.3 (1.9) | 16.8(3.3) | 26.6(3.6) |
| | 0 to <1x poverty line | 876 | 45.5(1.9) | 17.6(1.6) | 13.1(1.4) | 8.0(1.3) | 5.6(0.8) | 10.2 (1.4) |
| | 1x to <2x poverty line | 856 | 42.1(2.2) | 17.0 (1.8) | 11.9 (1.7) | 8.7(1.0) | 9.4(1.2) | 10.9(1.4) |
| | 2x to <3x poverty line | 516 | 40.1(2.7) | 20.8(2.9) | 13.7(1.9) | 4.9(0.8) | 8.5(1.5) | 12.0 (1.8) |
| Finfish Only | 3x to <4x poverty line | 368 | 37.0(3.9) | 16.1(2.6) | 11.1(1.9) | 9.7(2.1) | 11.8 (2.5) | 14.3(3.0) |
| | 4x to <5x poverty line | 251 | 32.0 (3.8) | 15.1(3.2) | 15.5 (3.4) | 8.7(1.9) | 11.1(3.0) | 17.5 (3.2) |
| | >= 5x poverty line | 498 | 31.9(2.7) | 14.1(2.1) | 12.1(2.0) | 10.6(1.6) | 12.4 (2.1) | 18.9 (1.8) |
| | Missing/Refused/DK | 343 | 34.3 (3.2) | 18.3 (2.3) | 17.4 (3.2) | 6.3 (1.5) | 12.6 (2.5) | 11.0(2.3) |
| | 0 to <1x poverty line | 876 | 56.5(2.4) | 16.6(1.3) | 9.2(1.2) | 7.0 (1.1) | 6.0(1.0) | 4.6(0.8) |
| | 1x to <2x poverty line | 856 | 54.2(2.7) | 18.2 (1.8) | 10.1(1.2) | 5.0(0.9) | 5.9(0.8) | 6.6(1.3) |
| | 2x to <3x poverty line | 516 | 49.4 (3.1) | 16.5 (2.6) | 14.0 (2.1) | 6.6 (1.5) | 6.2(1.3) | 7.2(1.4) |
| Shellfish Only | 3x to <4x poverty line | 368 | 48.4 (4.5) | 17.5 (3.0) | 15.8 (2.5) | 3.2(0.9) | 7.9 (2.1) | 7.2(2.1) |
| | 4x to <5x poverty line | 251 | 44.0 (4.3) | 19.5 (3.6) | 9.9(1.9) | 8.5(2.7) | 9.4 (2.2) | 8.6(2.3) |
| | >= 5x poverty line | 498 | 39.6 (3.5) | 16.0 (2.3) | 13.8(2.0) | 7.4(1.3) | 9.8 (2.3) | 13.4 (1.9) |
| | Missing/Refused/DK | 343 | 41.4 (3.9) | 18.1(2.6) | 12.5(2.4) | 14.4 (2.9) | 4.7(1.1) | 8.8(2.1) |

Table A.4a.Weighted percentages and their standard errors for categorized reports of 30-day frequency of consumption of fish, by NHANES survey
releases, income, race/ethnicity, age, and education, women aged 16-49 years, NHANES 2013-March 2020

| Devenueter | | | | | Percent (standard | l error) | | |
|----------------|--------------------|-------|------------|------------|-------------------|------------|------------|------------|
| Parameter | | N | 0 times | 1 time | 2 times | 3 times | 4-5 times | 6+ times |
| Race/Ethnicity | | | | | | | | |
| | Mexican American | 624 | 23.4(1.8) | 16.9(1.7) | 14.6(1.5) | 10.8(1.4) | 14.5(1.5) | 19.9(2.1) |
| | Other Hispanic | 380 | 21.2(2.3) | 17.8 (2.1) | 12.4(2.9) | 9.6(1.4) | 12.7(2.8) | 26.3(2.9) |
| Total Fish | Non-Hispanic White | 1,199 | 30.0(1.7) | 15.4 (1.2) | 10.4(1.3) | 7.5(0.9) | 11.8 (1.3) | 24.8(1.7) |
| | Non-Hispanic Black | 877 | 18.4 (2.1) | 12.5(1.1) | 13.4 (1.3) | 10.7(1.2) | 15.6(1.4) | 29.4(2.5) |
| | Other Race | 628 | 22.3 (2.1) | 10.1(1.5) | 11.0 (1.3) | 9.2(1.6) | 12.1(2.1) | 35.4(2.7) |
| | Mexican American | 624 | 40.5(2.2) | 20.2 (1.7) | 14.5(1.3) | 8.9(1.3) | 6.0(1.2) | 9.9(1.9) |
| | Other Hispanic | 380 | 34.5(2.3) | 20.5 (1.9) | 13.8(2.0) | 7.8 (1.7) | 8.9(1.3) | 14.5 (2.0) |
| Finfish Only | Non-Hispanic White | 1,199 | 41.4 (1.5) | 16.0(1.3) | 12.2(1.1) | 7.2(0.8) | 10.5 (1.1) | 12.7(1.0) |
| , | Non-Hispanic Black | 877 | 31.8 (2.2) | 16.7(1.8) | 16.7(1.4) | 9.8(1.4) | 10.1(1.2) | 14.8 (2.0) |
| | Other Race | 628 | 29.2(2.2) | 15.7(1.5) | 10.9(1.3) | 12.6 (2.2) | 11.2 (1.6) | 20.3 (2.2) |
| | Mexican American | 624 | 39.8(2.2) | 24.5 (1.6) | 15.4 (1.4) | 8.3 (1.3) | 7.7(1.0) | 4.2(1.0) |
| | Other Hispanic | 380 | 43.9(2.7) | 21.6 (2.3) | 12.0 (1.7) | 6.7(1.3) | 7.1(1.5) | 8.6(2.2) |
| Shellfish Only | Non-Hispanic White | 1,199 | 53.1(2.9) | 16.2(1.4) | 10.7(1.0) | 6.3 (1.1) | 6.9(1.2) | 6.7(0.8) |
| , | Non-Hispanic Black | 877 | 42.5 (2.7) | 13.6(1.3) | 15.8(1.4) | 8.3 (1.1) | 8.6 (1.2) | 11.3 (1.5) |
| | Other Race | 628 | 43.6(2.9) | 16.0 (1.9) | 11.4 (1.6) | 6.5 (1.2) | 6.5 (0.9) | 15.9(1.9) |
| Age | | | | | | | | |
| | 16 to 19 years | 654 | 41.1(2.6) | 18.3 (2.6) | 11.2 (1.5) | 8.6(1.2) | 10.3(2.0) | 10.5 (1.7) |
| | 20 to 29 years | 963 | 29.2 (1.8) | 14.5(1.4) | 9.8(1.1) | 8.7(1.2) | 12.0(1.3) | 25.8 (1.8) |
| Total Fish | 30 to 39 years | 1,056 | 24.4 (1.7) | 13.4 (1.2) | 10.8 (1.5) | 8.6 (1.0) | 12.2 (1.1) | 30.6 (1.9) |
| | 40 to 49 years | 1,035 | 19.3 (1.7) | 15.2 (1.7) | 14.1(1.5) | 8.8 (1.0) | 15.0(1.4) | 27.7 (2.0) |
| | 16 to 19 years | 654 | 58.2(2.6) | 19.6 (2.1) | 8.3(1.2) | 7.5 (1.5) | 3.7(0.9) | 2.6(0.6) |
| | 20 to 29 years | 963 | 42.1(1.8) | 14.9(1.6) | 13.9(1.7) | 6.2(1.0) | 9.1(1.3) | 13.8 (1.3) |
| Finfish Only | 30 to 39 years | 1,056 | 33.9(1.8) | 16.1(1.4) | 14.0 (1.1) | 9.5 (1.7) | 11.9 (1.5) | 14.5(1.4) |
| | 40 to 49 years | 1,035 | 30.8(2.1) | 18.7 (1.9) | 13.0(1.4) | 9.7(1.3) | 11.1(1.5) | 16.7(1.5) |
| | 16 to 19 years | 654 | 59.4(2.6) | 15.9(1.9) | 8.5(1.0) | 7.2 (1.3) | 5.5(1.4) | 3.6(0.9) |
| | 20 to 29 years | 963 | 48.1(2.6) | 17.3 (1.8) | 11.5 (1.3) | 5.8(1.0) | 8.6(1.2) | 8.7(1.2) |
| Shellfish Only | 30 to 39 years | 1,056 | 46.3 (2.5) | 15.9(1.2) | 11.3 (1.2) | 8.3 (1.2) | 8.3 (1.3) | 9.9(1.2) |
| | 40 to 49 years | 1,035 | 46.8(2.7) | 19.2 (1.6) | 15.0(1.3) | 6.3 (1.1) | 5.4 (1.0) | 7.3 (1.2) |
| Education | | ., | | | | | | |
| | < Median for age | 1,175 | 33.8(2.0) | 16.9(1.6) | 12.3(1.3) | 9.1(1.0) | 10.9(1.3) | 16.9(1.5) |
| | Median for age | 1,256 | 24.3(1.7) | 16.4 (1.3) | 11.9(1.2) | 8.7(1.0) | 13.8(1.3) | 24.9(1.6) |
| Total Fish | >Median for age | 904 | 19.9(2.0) | 10.8 (1.3) | 10.7(1.4) | 8.0(1.2) | 12.9(1.4) | 37.7(2.6) |
| | Unknown education | 373 | 35.6(2.7) | 17.6 (3.3) | 8.9(1.6) | 9.9(1.7) | 15.1(3.6) | 13.0 (2.5) |
| | < Median for age | 1,175 | 46.4(2.0) | 19.1(1.7) | 13.8 (1.3) | 6.7(0.9) | 6.8(1.0) | 7.3 (0.9) |
| | Median for age | 1,256 | 36.5(1.7) | 18.1(1.4) | 13.3 (1.2) | 7.8(1.0) | 10.4(1.0) | 14.0 (1.2) |
| Finfish Only | >Median for age | 904 | 30.0 (2.1) | 13.3 (1.5) | 13.0 (1.4) | 9.9 (1.1) | 13.2(1.6) | 20.5(1.7) |
| | Unknown education | 373 | 55.2(3.6) | 18.8 (2.9) | 7.4 (1.6) | 11.7 (2.5) | 3.3 (1.1) | 3.5(1.0) |

| Davamatar | | | | | Percent (standard | l error) | | |
|-----------|-------------------|-------|------------|------------|-------------------|-----------|------------|------------|
| Parameter | | N | 0 times | 1 time | 2 times | 3 times | 4-5 times | 6+ times |
| | < Median for age | 1,175 | 55.0(2.5) | 17.7(1.4) | 12.1(1.3) | 5.8(0.9) | 4.6(0.9) | 4.8(0.8) |
| 01-116-1 | Median for age | 1,256 | 49.7(2.6) | 17.7(1.5) | 11.0 (0.9) | 6.9(0.8) | 6.6(0.7) | 8.0 (1.1) |
| Shellfish | >Median for age | 904 | 40.6(2.8) | 16.2(1.7) | 13.8(1.4) | 7.4(1.4) | 10.3 (1.7) | 11.7 (1.4) |
| | Unknown education | 373 | 51.6 (3.3) | 18.6 (2.2) | 8.6(1.4) | 9.0 (1.8) | 6.8 (2.1) | 5.4(1.8) |

Note: Data from NHANES 2011-2012 data was not included in the current study.

| Darameter | | N | | | Percent (sta | ndard error) | | |
|----------------|-----------------------|-------|------------|------------|--------------|--------------|------------|------------|
| Parameter | | N | 0 times | 1 time | 2 times | 3 times | 4-5 times | 6+ times |
| NHANES Survey | y Release | | | | | | | |
| | 1999-2000 | 1,637 | 21.8(2.3) | 14.7(1.1) | 13.6(1.2) | 8.9(1.6) | 14.6(1.0) | 26.4(3.0) |
| | 2001-2002 | 1,780 | 16.9(1.1) | 13.3(1.4) | 13.5(1.6) | 10.4(1.0) | 16.9(1.2) | 28.9(1.5) |
| Total Fish | 2003-2004 | 1,599 | 21.3 (1.6) | 12.5(1.7) | 13.0(1.1) | 9.8(0.9) | 14.4(1.2) | 29.0(2.0) |
| TULUI FISII | 2005-2006 | 1,792 | 20.1(2.2) | 10.7(1.1) | 11.9 (0.8) | 9.4(0.9) | 13.8 (1.1) | 34.2(3.0) |
| | 2007-2008 | 1,493 | 24.7(1.5) | 14.8(1.0) | 11.4 (0.7) | 9.4(0.9) | 12.9(1.0) | 26.9(1.5) |
| | 2009-2010 | 1,786 | 20.2(1.9) | 14.5(0.9) | 10.5(0.6) | 7.8(0.8) | 15.2 (1.1) | 31.8(1.6) |
| | 1999-2000 | 1,637 | 32.9(1.8) | 17.7(1.1) | 16.9(0.8) | 7.5 (1.9) | 11.4 (1.2) | 13.6(1.9) |
| | 2001-2002 | 1,780 | 25.6(1.5) | 17.2 (1.4) | 17.0 (1.5) | 11.5 (0.7) | 13.4(1.4) | 15.4 (1.3) |
| Firfich Orly | 2003-2004 | 1,599 | 31.2 (2.1) | 16.8 (1.3) | 14.6(0.9) | 10.3(1.0) | 12.0 (1.1) | 15.2(1.4) |
| Finfish Only | 2005-2006 | 1,792 | 29.8(2.6) | 14.6(1.0) | 12.8(0.7) | 10.1(0.7) | 15.3 (0.9) | 17.4 (2.5) |
| | 2007-2008 | 1,493 | 35.9(1.6) | 16.0(1.2) | 13.5(0.9) | 9.1(0.7) | 11.6 (1.0) | 14.0(1.3) |
| | 2009-2010 | 1,786 | 32.5(1.9) | 15.1(1.0) | 14.3(1.2) | 8.6(1.1) | 13.3(1.3) | 16.1(1.2) |
| | 1999-2000 | 1,637 | 48.3 (3.3) | 19.4 (1.8) | 9.4 (1.3) | 7.0 (1.0) | 9.4 (1.8) | 6.5(1.7) |
| | 2001-2002 | 1,780 | 48.3(1.7) | 17.9 (1.7) | 11.8 (0.9) | 7.8(0.8) | 5.6(0.9) | 8.7(1.1) |
| | 2003-2004 | 1,599 | 48.9(2.4) | 16.5(1.1) | 11.0 (1.2) | 8.6(1.1) | 7.8(0.9) | 7.2 (1.0) |
| Shellfish Only | 2005-2006 | 1,792 | 44.4(2.7) | 17.1(1.7) | 12.9(1.5) | 6.9(0.8) | 9.3 (1.0) | 9.5 (1.1) |
| | 2007-2008 | 1,493 | 49.9(1.5) | 17.0 (1.0) | 11.1(1.0) | 6.9(0.9) | 7.3(0.9) | 7.8(0.7) |
| | 2009-2010 | 1,786 | 42.2(1.9) | 17.0 (0.9) | 13.7(1.0) | 9.2(0.8) | 8.2 (1.0) | 9.7(1.5) |
| Income | · | | | | | | | |
| | <\$20,000 | 2,216 | 24.8(1.5) | 16.1(0.9) | 14.5(1.1) | 9.8 (1.1) | 11.8 (0.8) | 23.0(1.4) |
| | \$20,000 to<\$45,000 | 2,894 | 22.9(1.4) | 14.8(0.9) | 12.4 (0.9) | 9.5(0.7) | 14.4 (0.8) | 26.0(1.2) |
| | \$45,000 to <\$75,000 | 1,950 | 21.6 (1.5) | 13.7(0.9) | 11.0(0.8) | 9.7(0.8) | 14.7(1.0) | 29.4 (1.6) |
| Total Fish | \$75,000 and over | 2,148 | 16.2 (1.0) | 10.3(0.8) | 12.2(0.8) | 8.5(0.8) | 16.4 (0.9) | 36.3 (1.5) |
| | \$20,000 and over | 225 | 21.3(3.6) | 9.9(2.5) | 10.8(3.3) | 12.9(2.6) | 9.6(2.4) | 35.5(3.8) |
| | Refused/Don't Know | 163 | 20.7(3.9) | 20.9(4.4) | 13.3(3.6) | 9.0(2.8) | 9.6(2.9) | 26.4(5.8) |
| | Uncalculated* | 491 | 21.5 (2.5) | 15.2(2.2) | 10.7(2.3) | 7.4 (1.7) | 17.7(2.7) | 27.4(3.8) |

Table A.4b.Percentages and their standard errors for categorized reports of 30-day frequency of consumption of fish, by NHANES survey release,
income, race/ethnicity, and age for women aged 16-49 years, NHANES 1999-2010 (United States Environmental Protection Agency, 2013)

| Davamenter | | N | | | Percent (sta | ndard error) | | |
|-----------------------|-----------------------|-------|------------|------------|--------------|--------------|------------|------------|
| Parameter | | N | 0 times | 1 time | 2 times | 3 times | 4-5 times | 6+ times |
| | <\$20,000 | 2,216 | 35.7(1.7) | 19.7(1.0) | 14.9(1.0) | 7.5(0.9) | 10.7(1.0) | 11.6(1.3) |
| | \$20,000 to<\$45,000 | 2,894 | 33.6(1.6) | 18.0 (1.0) | 15.4(1.0) | 8.6(0.6) | 10.9(0.8) | 13.5 (1.1) |
| | \$45,000 to <\$75,000 | 1,950 | 31.8 (1.6) | 16.2 (1.1) | 15.5(1.1) | 9.2(0.8) | 11.8 (1.0) | 15.5(1.3) |
| Finfish Only | \$75,000 and over | 2,148 | 26.2(1.2) | 12.7(0.8) | 14.7(0.9) | 11.1(0.8) | 16.9 (1.1) | 18.4 (1.1) |
| - | \$20,000 and over | 225 | 31.1(3.5) | 10.6(2.3) | 11.8 (3.2) | 12.9(2.5) | 10.7(2.8) | 23.0(3.3) |
| | Refused/Don't Know | 163 | 39.0 (6.8) | 25.4 (4.8) | 9.0(2.2) | 5.2 (2.1) | 7.3(2.4) | 14.2(5.2) |
| | Uncalculated* | 491 | 32.7(3.2) | 16.6(2.0) | 12.0(2.0) | 12.6(3.4) | 12.2(2.5) | 13.9(2.7) |
| | <\$20,000 | 2,216 | 55.2(1.4) | 16.2 (1.1) | 9.8(0.8) | 6.0(0.7) | 5.9(0.7) | 6.9(0.8) |
| | \$20,000 to<\$45,000 | 2,894 | 50.3(1.6) | 16.7(0.9) | 10.4 (0.8) | 7.7(0.7) | 6.9(0.8) | 8.0(0.7) |
| | \$45,000 to <\$75,000 | 1,950 | 45.8(1.8) | 17.5(1.0) | 12.8(1.0) | 8.0(0.8) | 8.8(1.0) | 7.1(0.9) |
| Shellfish Only | \$75,000 and over | 2,148 | 40.4(1.5) | 18.3 (1.2) | 13.5(0.9) | 8.8(0.7) | 8.9(1.0) | 10.1(0.9) |
| - | \$20,000 and over | 225 | 46.6(4.5) | 16.8(2.7) | 10.4 (2.7) | 7.7(2.2) | 8.9(2.7) | 9.7(2.9) |
| | Refused/Don't Know | 163 | 41.0 (4.9) | 24.8(4.4) | 9.1(2.9) | 5.4 (1.8) | 15.6(4.1) | 4.0(1.6) |
| | Uncalculated* | 491 | 49.2 (3.4) | 19.7(3.0) | 8.9(1.6) | 6.1(1.7) | 7.4 (1.8) | 8.6(2.3) |
| Race/Ethnicity | | | | | | | | |
| | Mexican American | 2,589 | 22.9(1.1) | 16.9(1.0) | 15.9(1.0) | 11.8(0.8) | 14.2(0.9) | 18.3 (1.1) |
| | Other Hispanic | 751 | 21.1(2.3) | 20.3 (1.9) | 10.2(1.6) | 9.7(1.2) | 12.3(1.5) | 26.5(2.8) |
| Total Fish | Non-Hispanic White | 4,043 | 21.4 (0.9) | 12.9(0.7) | 12.3 (0.6) | 9.2(0.6) | 14.4 (0.6) | 29.8(1.2) |
| | Non-Hispanic Black | 2,230 | 17.1(1.2) | 12.9(0.8) | 13.0(0.8) | 8.6(0.7) | 18.2 (1.1) | 30.2(1.2) |
| | Other Race | 474 | 19.5(2.6) | 7.1(1.2) | 7.2(1.2) | 7.0(1.4) | 12.6(1.8) | 46.6(3.4) |
| | Mexican American | 2,589 | 38.9(1.5) | 21.4 (1.0) | 15.0(0.8) | 8.7(0.8) | 8.3(0.6) | 7.8(0.8) |
| | Other Hispanic | 751 | 34.7(2.6) | 20.9(2.4) | 13.9(1.5) | 7.1(1.1) | 8.4(1.2) | 15.0(2.2) |
| Finfish Only | Non-Hispanic White | 4,043 | 31.5 (1.0) | 14.9(0.7) | 15.1(0.6) | 9.6(0.6) | 13.4 (0.7) | 15.5(0.8) |
| - | Non-Hispanic Black | 2,230 | 25.8(1.3) | 19.0 (1.1) | 15.7(0.9) | 10.3(0.8) | 13.6(0.9) | 15.6(0.9) |
| | Other Race | 474 | 25.7(2.6) | 11.9 (1.9) | 10.7(1.6) | 10.3(1.4) | 17.2 (1.9) | 24.1(3.0) |
| | Mexican American | 2,589 | 44.7(1.6) | 21.6 (1.1) | 14.2 (0.9) | 8.2(0.5) | 6.5(0.7) | 4.7(0.5) |
| | Other Hispanic | 751 | 47.5(3.0) | 19.5 (1.7) | 10.5(1.4) | 9.1(1.2) | 5.8 (1.1) | 7.7(1.2) |
| Shellfish Only | Non-Hispanic White | 4,043 | 48.3(1.2) | 17.5(0.8) | 11.3 (0.6) | 7.2(0.5) | 7.9(0.6) | 7.7(0.7) |
| - | Non-Hispanic Black | 2,230 | 45.5(1.4) | 16.3 (1.1) | 12.7(0.8) | 7.4(0.7) | 8.4(0.6) | 9.6(0.8) |
| | Other Race | 474 | 38.4 (3.1) | 10.6(1.7) | 10.6(1.6) | 11.8 (1.8) | 11.6 (1.5) | 17.0 (2.2) |

| Parameter | | N | | | Percent (sta | ndard error) | | |
|----------------|----------------|-------|------------|------------|--------------|--------------|------------|------------|
| Parameter | | N | 0 times | 1 time | 2 times | 3 times | 4-5 times | 6+ times |
| Age | | | | | | | | |
| | 16 to 19 years | 2,439 | 38.9(1.6) | 17.9(1.2) | 10.4 (0.7) | 7.2(0.7) | 10.6(0.9) | 15.1(1.1) |
| Tatal | 20 to 29 years | 2,739 | 23.4(1.3) | 14.2(1.0) | 13.2(0.9) | 9.8(0.9) | 13.6(1.0) | 25.7(1.4) |
| Total Fish | 30 to 39 years | 2,495 | 18.2 (1.1) | 12.5(1.0) | 12.4 (0.9) | 9.8(0.7) | 14.6(0.8) | 32.4(1.4) |
| | 40 to 49 years | 2,414 | 1.05 (1.0) | 12.1(0.8) | 12.0 (0.7) | 9.0(0.7) | 16.8 (0.8) | 35.1(1.3) |
| | 16 to 19 years | 2,439 | 53.7(1.7) | 16.5(1.1) | 9.3(0.8) | 6.2(0.7) | 6.9(0.8) | 7.4(0.8) |
| Finfich Only | 20 to 29 years | 2,739 | 36.1(1.4) | 16.9(1.0) | 13.7(0.9) | 8.9(0.7) | 11.3 (0.8) | 13.1(1.1) |
| Finfish Only | 30 to 39 years | 2,495 | 28.7(1.2) | 16.1(0.9) | 15.4 (0.9) | 10.3 (0.7) | 13.2 (0.8) | 16.2 (0.9) |
| | 40 to 49 years | 2,414 | 22.1(1.1) | 15.6(0.9) | 17.2 (1.0) | 10.4 (0.8) | 15.8 (0.9) | 19.0(1.2) |
| | 16 to 19 years | 2,439 | 58.8(1.4) | 17.9(1.3) | 9.4 (0.8) | 4.7(0.5) | 4.9(0.5) | 4.3(0.7) |
| Challfich Only | 20 to 29 years | 2,739 | 47.9(1.4) | 18.2 (1.0) | 11.5 (0.8) | 7.6(0.7) | 7.4(0.7) | 7.4(0.7) |
| Shellfish Only | 30 to 39 years | 2,495 | 43.7(1.5) | 18.1(1.0) | 11.4 (0.8) | 8.2(0.7) | 8.8(0.8) | 9.7(0.8) |
| | 40 to 49 years | 2,414 | 45.3(1.5) | 16.0 (1.0) | 12.7(0.9) | 8.4(0.7) | 8.6(0.9) | 8.9(0.8) |

* Uncalculated indicates that the participant is residing in a multi-family dwelling and one or more of the families only reported a range for their family income, either <\$20,000 or >\$20,000. Thus NCHS did not calculate household income for these participants⁻

| | | N | Arithmetic mean | | | Selected percentiles | (95% CI) | |
|--------------------|--------------------|----------|-----------------------|-----------------|--------------------|----------------------|--------------------|-----------------------|
| | | N | (95% CI) | 25th | 50th | 75th | 90th | 95th |
| Estimated <i>i</i> | Amount of Fish (| Consumed | l (g) in Last 30 Days | | | | | |
| | 2013-2014 | 814 | 97.6(69.2,126.0) | 0.0(-20.7,20.7) | 28.1(7.3,48.9) | 107.5(74.0,141.1) | 266.7(186.9,346.5) | 407.4(285.6,529.1) |
| Shellfish | 2015-2016 | 740 | 91.3(77.8,104.7) | 0.0(-21.2,21.2) | 28.6(7.4,49.8) | 99.3(70.4,128.2) | 271.0(189.8,352.3) | 417.4(354.2,480.7) |
| Sheimsh | 2017-March 2020 | 2154 | 93.0(79.8,106.1) | 0.0(-20.9,20.9) | 37.9(16.7,59.0) | 104.6(76.3,132.9) | 249.3(201.1,297.5) | 393.7(326.9,460.4) |
| | 2013-2014 | 814 | 237.8(203.7,271.8) | 0.0(-34.4,34.4) | 81.4(57.5,105.3) | 301.0(234.7,367.3) | 679.8(585.7,773.9) | 952.5(706.5,1198.6) |
| Finfish | 2015-2016 | 740 | 227.5(190.6,264.4) | 0.0(-34.4,34.4) | 98.4(67.7,129.0) | 299.4(237.3,361.4) | 597.4(455.1,739.6) | 918.6(671.1,1166.0) |
| FINIISN | 2017-March 2020 | 2154 | 232.6(212.8,252.4) | 0.0(-30.6,30.6) | 102.7(78.0,127.4) | 296.8(259.1,334.4) | 643.9(555.5,732.3) | 937.7(771.5,1103.9) |
| | 2013-2014 | 814 | 335.3(278.9,391.8) | 0.0(-27.9,27.9) | 143.2(92.3,194.0) | 435.4(329.9,540.9) | 855.3(739.2,971.3) | 1216.6(911.8,1521.5) |
| Total Fish | 2015-2016 | 740 | 318.8(271.4,366.2) | 0.0(-22.3,22.3) | 161.1(111.5,210.7) | 413.8(327.9,499.6) | 789.1(616.8,961.3) | 1114.6(811.3,1418.0) |
| i utui FISH | 2017-March 2020 | 2154 | 325.6(296.2,354.9) | 0.0(-21.3,21.3) | 161.5(127.7,195.2) | 420.4(364.2,476.7) | 822.1(716.7,927.4) | 1228.8(1077.5,1380.0) |

Table A.5. Estimated amount of fish consumed (g) in last 30 days, by NHANES survey releases, women aged 16-49 years, NHANES 2013-March 2020

| | | N | Arithmetic mean | | | Selected percentiles (| 95% CI) | |
|------------|------------------------|-----------|--------------------|------------------|------------------|------------------------|--------------------|----------------------|
| | | N | (95% CI) | 25th | 50th | 75th | 90th | 95th |
| Estimated | Mercury Intake (ug) in | n Last 30 | Days | | | | | |
| | 2013-2014 | 814 | 3.65(2.30,4.99) | 0.00(-0.29,0.29) | 0.34(0.05,0.63) | 2.71(1.02,4.39) | 10.01(4.57,15.46) | 17.51(13.26,21.77) |
| Shellfish | 2015-2016 | 740 | 3.37(2.75,3.99) | 0.00(-0.29,0.29) | 0.36(0.07,0.65) | 2.05(0.56,3.55) | 11.87(7.72,16.02) | 19.47(13.46,25.48) |
| | 2017-March 2020 | 2154 | 3.22(2.74,3.70) | 0.00(-0.29,0.29) | 0.52(0.23,0.81) | 2.10(1.02,3.18) | 9.63(6.69,12.56) | 15.84(13.93,17.75) |
| | 2013-2014 | 814 | 21.97(19.47,24.47) | 0.00(-1.82,1.82) | 8.03(5.29,10.78) | 26.25(22.40,30.09) | 57.05(46.69,67.41) | 91.87(68.70,115.04) |
| Finfish | 2015-2016 | 740 | 19.46(16.42,22.50) | 0.00(-1.76,1.76) | 8.55(6.99,10.12) | 23.04(17.89,28.20) | 47.94(35.97,59.90) | 81.46(50.33,112.59) |
| | 2017-March 2020 | 2154 | 20.51(18.37,22.65) | 0.00(-1.70,1.70) | 8.02(6.85,9.19) | 23.63(20.57,26.69) | 54.35(46.25,62.44) | 85.81(69.88,101.74) |
| | 2013-2014 | 814 | 25.62(22.16,29.08) | 0.00(-0.57,0.57) | 9.39(6.48,12.30) | 30.93(26.22,35.64) | 66.46(51.48,81.44) | 104.26(75.57,132.94) |
| Total Fish | 2015-2016 | 740 | 22.83(19.55,26.12) | 0.00(-0.30,0.30) | 9.91(7.34,12.47) | 29.16(24.93,33.39) | 55.94(39.64,72.23) | 84.27(51.47,117.07) |
| | 2017-March 2020 | 2154 | 23.73(21.24,26.21) | 0.00(-0.29,0.29) | 9.55(7.94,11.16) | 28.59(25.14,32.03) | 62.66(55.21,70.11) | 91.02(75.33,106.72) |

Table A.6. Estimated mercury intake (µg) in last 30 days, by NHANES survey releases, women aged 16-49 years, NHANES 2013-March 2020

| | ITTAITEO 2010 TH | | | | | | | | |
|------------|------------------------|------------|------------------------|-------------------------------|-----------------|-----------------|-----------------|-----------------|--|
| | | N | Arithmetic Mean | Selected percentiles (95% CI) | | | | | |
| | | N | (95% CI) | 25th | 50th | 75th | 90th | 95th | |
| Estimated | Mercury Intake per Uni | t Body Wei | ght (µg Hg/kg bw) in L | .ast 30 Days | | | | | |
| | 2013-2014 | 814 | 0.05(0.03,0.06) | 0.00(0.00,0.00) | 0.00(0.00,0.01) | 0.03(0.01,0.05) | 0.13(0.07,0.19) | 0.24(0.16,0.31) | |
| Shellfish | 2015-2016 | 740 | 0.05(0.04,0.06) | 0.00(0.00,0.00) | 0.00(0.00,0.01) | 0.03(0.02,0.05) | 0.15(0.10,0.20) | 0.26(0.21,0.32) | |
| | 2017-March 2020 | 2154 | 0.04(0.04,0.05) | 0.00(0.00,0.00) | 0.01(0.00,0.01) | 0.03(0.02,0.05) | 0.12(0.09,0.15) | 0.22(0.19,0.25) | |
| | 2013-2014 | 814 | 0.31(0.28,0.35) | 0.00(-0.03,0.03) | 0.09(0.05,0.13) | 0.36(0.30,0.42) | 0.84(0.60,1.08) | 1.30(0.98,1.62) | |
| Finfish | 2015-2016 | 740 | 0.27(0.23,0.31) | 0.00(-0.03,0.03) | 0.10(0.07,0.12) | 0.32(0.26,0.38) | 0.72(0.53,0.91) | 1.12(0.78,1.45) | |
| | 2017-March 2020 | 2154 | 0.28(0.25,0.31) | 0.00(-0.02,0.02) | 0.10(0.08,0.12) | 0.32(0.27,0.37) | 0.74(0.60,0.87) | 1.15(0.90,1.41) | |
| Total Fish | 2013-2014 | 814 | 0.36(0.31,0.41) | 0.00(-0.01,0.01) | 0.13(0.08,0.17) | 0.44(0.35,0.52) | 0.93(0.69,1.17) | 1.39(0.92,1.87) | |
| | 2015-2016 | 740 | 0.32(0.27,0.37) | 0.00(0.00,0.00) | 0.13(0.10,0.17) | 0.39(0.32,0.46) | 0.82(0.60,1.04) | 1.34(0.92,1.76) | |
| | 2017-March 2020 | 2154 | 0.32(0.29,0.36) | 0.00(0.00,0.00) | 0.13(0.10,0.16) | 0.38(0.34,0.43) | 0.81(0.71,0.92) | 1.30(1.09,1.51) | |

Table A.7.Estimated mercury intake per unit body weight (µg Hg/kg bw) in last 30 days, by NHANES survey releases, women aged 16-49 years,
NHANES 2013-March 2020

| Parameter | | N | Arithmetic mean (95% Cl) | 25th | 50th | 75th | 90th | 95th |
|--------------------|--|------|-----------------------------|------------------|--------------------|--------------------|----------------------|-----------------------|
| Estimated An | nount of Fish Consumed (g) | | | | | | | |
| D (| Mexican American | 624 | 300.4(250.9,349.9 | 37.6(17.7,57.5) | 149.3(126.3,172.3) | 373.1(308.8,437.3) | 706.0(549.8,862.3) | 1013.3(709.0,1317.5) |
| | Other Hispanic | 380 | 329.6(267.9,391.3) | 32.6(14.7,50.5) | 155.2(112.3,198.2) | 424.1(322.0,526.3 | 769.2(544.4,994.0) | 1331.9(678.9,1984.8) |
| Race/ Ethnicity | Non-Hispanic White | 1199 | 305.4(278.1,332.7) | 0.0(-20.5,20.5) | 143.4(109.4,177.4) | 412.7(352.5,473.0) | 791.8(715.6,868.0) | 1151.5(972.1,1330.9) |
| Limicity | Non-Hispanic Black | 877 | 443.2(375.9,510.5) | 71.2(38.7,103.6) | 232.0(196.2,267.8 | 510.9(424.2,597.7) | 1057.1(764.7,1349.5) | 1600.3(1048.8,2151.8) |
| | Other Race | 628 | 324.7(288.3,361.1) | 38.8(14.8,62.7) | 174.7(130.6,218.7) | 449.0(387.1,510.9) | 871.1(694.4,1047.8) | 1170.1(973.9,1366.4) |
| | 16 to 19 | 654 | 142.0(122.8,161.1) | 0.0(-21.2,21.2) | 59.5(42.8,76.2) | 200.6(157.4,243.9) | 384.4(295.2,473.6) | 594.0(496.3,691.8) |
| Age Vegre | 20 to 29 | 963 | 319.4(286.1,352.7) | 0.0(-21.4,21.4) | 147.8(118.9,176.7) | 436.9(374.3,499.4) | 843.9(711.4,976.3) | 1146.5(977.5,1315.5) |
| Age, Years | 30 to 39 | 1056 | 374.6(336.3,413.0) | 31.4(9.0,53.8) | 186.9(138.7,235.0) | 503.0(430.1,575.9) | 959.7(800.7,1118.7) | 1300.1(1052.0,1548.1) |
| | 40 to 49 | 1035 | 358.5(326.7,390.3) | 58.9(44.0,73.8) | 195.5(158.4,232.5) | 464.5(400.9,528.1) | 857.6(800.3,914.9) | 1391.3(1106.7,1675.9) |
| | 0 to <1x poverty line | 876 | 265.9(223.9,307.9) | 0.0(-21.1,21.1) | 105.3(78.4,132.2) | 317.7(254.0,381.4) | 689.2(556.7,821.7) | 993.5(738.6,1248.3) |
| | 1x to <2x poverty | 856 | 277.0(243.7,310.3) | 0.0(-21.4,21.4) | 130.8(97.5,164.2) | 389.2(339.6,438.8) | 734.6(618.8,850.3) | 1039.6(784.4,1294.9) |
| A | 2x to <3x poverty line | 516 | 295.4(250.0,340.8) | 0.0(-29.0,29.0) | 136.4(101.2,171.7) | 369.1(280.8,457.5) | 762.5(547.6,977.5) | 1184.7(813.2,1556.1) |
| Annual Income | 3x to <4x poverty line | 368 | 317.8(256.3,379.2) | 0.0(-33.6,33.6) | 161.4(97.9,224.9) | 463.5(314.6,612.4) | 820.2(631.9,1008.5) | 1104.4(717.8,1491.0) |
| liicome | 4x to <5x poverty line | 251 | 396.0(326.0,466.0) | 45.1(1.2,89.0) | 199.6(127.5,271.8) | 531.1(390.1,672.0) | 968.8(709.1,1228.5) | 1373.9(585.9,2161.8) |
| | >= 5x poverty line | 498 | 426.3(382.0,470.6) | 44.7(8.1,81.4) | 279.3(209.4,349.2) | 591.6(509.8,673.3) | 1085.6(954.2,1217.0) | 1349.8(1175.0,1524.6) |
| | Missing/Refused/DK | 343 | 331.6(287.7,375.6) | 42.4(14.2,70.7) | 192.6(134.4,250.7) | 380.8(316.6,444.9) | 773.9(631.5,916.3) | 1098.7(566.2,1631.3) |
| | <median age<="" education="" for="" td=""><td>1175</td><td>234.0(208.8,259.2)</td><td>0.0(-22.0,22.0)</td><td>95.8(71.0,120.6)</td><td>296.7(238.8,354.5)</td><td>630.0(539.0,721.0)</td><td>907.8(749.0,1066.5)</td></median> | 1175 | 234.0(208.8,259.2) | 0.0(-22.0,22.0) | 95.8(71.0,120.6) | 296.7(238.8,354.5) | 630.0(539.0,721.0) | 907.8(749.0,1066.5) |
| Education | Median education for age | 1256 | 337.9(297.0,378.8) | 32.1(10.5,53.7) | 155.0(126.9,183.1) | 416.2(358.9,473.5) | 801.5(684.7,918.3) | 1362.7(1098.3,1627.2) |
| Education | >Median education for age | 904 | 425.4(387.7,463.2) | 64.8(41.9,87.8) | 273.8(214.3,333.3) | 626.2(561.2,691.3) | 1024.1(904.8,1143.4) | 1376.7(1180.9,1572.4) |
| | Unknown education | 373 | 170.2(137.0,203.4) | 0.0(-20.9,20.9) | 76.4(38.2,114.6) | 254.2(182.6,325.7) | 437.8(307.4,568.1) | 656.0(516.1,796.0) |
| Estimated Me | rcury Intake (µg) | | | | | | | |
| | Mexican American | 624 | 22.30(17.24,27.37) | 0.43(0.19,0.68) | 9.39(6.61,12.18) | 25.28(19.43,31.12) | 56.13(47.26,65.01) | 82.21(43.82,120.60) |
| _ / | Other Hispanic | 380 | 24.49(19.88,29.10) | 0.41(0.17,0.65) | 10.39(7.56,13.23) | 28.28(21.98,34.57) | 63.97(52.32,75.62) | 83.77(50.14,117.41) |
| Race/ Ethnicity | Non-Hispanic White | 1199 | 22.83(20.90,24.75) | 0.00(-0.27,0.27) | 9.11(7.29,10.93) | 28.19(24.99,31.38) | 60.57(52.58,68.57) | 95.47(78.19,112.75) |
| | Non-Hispanic Black | 877 | 32.03(27.29,36.77) | 1.12(-0.29,2.52) | 15.38(11.82,18.95) | 36.45(31.17,41.73) | 72.63(56.01,89.26) | 123.50(87.85,159.15) |
| | Other Race | 628 | 22.15(19.27,25.04) | 0.86(-0.35,2.07) | 9.38(7.02,11.73) | 29.42(24.25,34.59) | 62.97(49.79,76.15) | 89.73(67.39,112.06) |
| | 16 to 19 | 654 | 9.12(7.94,10.31) | 0.00(-0.30,0.30) | 1.91(0.51,3.31) | 12.41(9.90,14.91) | 26.18(22.06,30.30) | 35.65(25.90,45.40) |
| Age, Years | 20 to 29 | 963 | 22.18(19.86,24.51) | 0.00(-0.30,0.30) | 9.14(7.50,10.78) | 27.23(23.90,30.56) | 63.87(52.50,75.23) | 88.93(71.51,106.35) |
| | 30 to 39 | 1056 | 26.69(23.82,29.56) | 0.40(-0.17,0.97) | 12.20(9.44,14.95) | 34.90(30.45,39.35) | 62.99(54.08,71.90) | 97.75(70.58,124.91) |
| | 40 to 49 | 1035 | 29.16(26.05,32.26) | 0.94(0.14,1.74) | 12.23(10.29,14.16) | 33.68(28.68,38.68) | 73.81(59.19,88.42) | 128.91(98.42,159.40) |

Table A.8. Estimated amounts of fish consumed (g), mercury intake (µg), and mercury intake per unit body weight (µg/kg), by race/ethnicity, age, income, and education, women aged 16-49 years, NHANES 2013-March 2020

| Parameter | | N | Arithmetic mean (95% Cl) | 25th | 50th | 75th | 90th | 95th |
|--------------------|---|------------|-----------------------------|------------------|--------------------|--------------------|---------------------|-----------------------|
| | 0 to <1x poverty line | 876 | 19.86(16.85,22.86) | 0.00(-0.30,0.30) | 5.09(2.39,7.80) | 21.05(17.57,24.52) | 53.88(43.83,63.92) | 77.27(56.77,97.77) |
| | 1x to <2x poverty | 856 | 20.93(18.40,23.46) | 0.00(-0.30,0.30) | 9.13(6.45,11.80) | 29.47(24.45,34.49) | 57.89(47.15,68.63) | 81.20(66.02,96.37) |
| | 2x to <3x poverty line | 516 | 20.88(17.23,24.53) | 0.00(-0.63,0.63) | 9.04(6.82,11.27) | 25.22(20.33,30.11) | 50.15(34.18,66.11) | 89.64(54.93,124.34) |
| Annual Income | 3x to <4x poverty line | 368 | 24.51(18.35,30.66) | 0.00(-0.78,0.78) | 11.17(6.59,15.74) | 29.99(20.99,38.99) | 55.44(31.80,79.08) | 90.28(33.13,147.43) |
| IIICOIIIE | 4x to <5x poverty line | 251 | 26.63(21.46,31.81) | 0.73(-0.87,2.32) | 12.32(8.30,16.34) | 33.69(25.86,41.52) | 68.71(31.54,105.89) | 113.20(55.38,171.01) |
| | >= 5x poverty line | 498 | 31.51(27.44,35.58) | 0.90(-0.70,2.51) | 16.96(12.44,21.49) | 38.01(32.23,43.79) | 80.89(63.29,98.49) | 113.00(82.33,143.68) |
| | Missing/Refused/DK | 343 | 24.78(19.75,29.81) | 0.59(-0.01,1.19) | 9.77(7.62,11.93) | 24.10(16.46,31.73) | 63.60(51.90,75.31) | 121.19(70.30,172.08) |
| | <median age<="" education="" for="" td=""><td>1175</td><td>17.79(15.82,19.76)</td><td>0.00(-0.30,0.30)</td><td>5.29(2.57,8.01)</td><td>19.82(16.25,23.39)</td><td>48.82(39.91,57.72)</td><td>72.54(61.62,83.45)</td></median> | 1175 | 17.79(15.82,19.76) | 0.00(-0.30,0.30) | 5.29(2.57,8.01) | 19.82(16.25,23.39) | 48.82(39.91,57.72) | 72.54(61.62,83.45) |
| Education | Median education for age | 1256 | 24.47(21.55,27.38) | 0.40(0.10,0.70) | 10.12(8.25,11.99) | 29.58(26.50,32.65) | 61.80(52.37,71.23) | 101.15(72.12,130.19) |
| Education | >Median education for age | 904 | 31.45(28.30,34.61) | 1.25(-0.09,2.60) | 16.17(12.27,20.06) | 38.70(34.86,42.54) | 76.40(62.63,90.17) | 129.84(100.67,159.02) |
| | Unknown education | 373 | 10.63(8.46,12.80) | 0.00(-0.30,0.30) | 2.60(0.78,4.42) | 13.30(8.43,18.18) | 28.74(24.80,32.68) | 44.81(27.84,61.78) |
| Estimated Me | ercury Intake per Unit Body We | ight (µg l | Hg/kg bw) | | | | | |
| | Mexican American | 624 | 0.32(0.24,0.41) | 0.01(0.00,0.01) | 0.12(0.09,0.16) | 0.34(0.27,0.42) | 0.79(0.66,0.92) | 1.20(0.75,1.66) |
| Race/ | Other Hispanic | 380 | 0.37(0.29,0.44) | 0.01(0.00,0.01) | 0.14(0.10,0.19) | 0.39(0.29,0.49) | 0.95(0.70,1.19) | 1.39(1.12,1.66) |
| Race/ Ethnicity | Non-Hispanic White | 1199 | 0.31(0.29,0.34) | 0.00(0.00,0.00) | 0.12(0.09,0.14) | 0.40(0.36,0.44) | 0.79(0.66,0.92) | 1.26(0.97,1.55) |
| Linneity | Non-Hispanic Black | 877 | 0.38(0.33,0.43) | 0.02(-0.01,0.04) | 0.18(0.14,0.22) | 0.44(0.40,0.49) | 0.96(0.68,1.24) | 1.42(1.13,1.72) |
| | Other Race | 628 | 0.36(0.31,0.41) | 0.01(-0.01,0.03) | 0.13(0.09,0.17) | 0.42(0.35,0.49) | 1.01(0.75,1.27) | 1.41(1.16,1.67) |
| | 16 to 19 | 654 | 0.14(0.12,0.16) | 0.00(-0.01,0.01) | 0.02(0.00,0.04) | 0.19(0.15,0.22) | 0.42(0.36,0.48) | 0.63(0.53,0.72) |
| Age, Years | 20 to 29 | 963 | 0.32(0.29,0.35) | 0.00(0.00,0.00) | 0.13(0.10,0.16) | 0.37(0.31,0.43) | 0.88(0.69,1.07) | 1.37(1.17,1.57) |
| Aye, reuis | 30 to 39 | 1056 | 0.36(0.31,0.40) | 0.00(0.00,0.01) | 0.15(0.11,0.19) | 0.45(0.40,0.49) | 0.88(0.75,1.02) | 1.32(0.96,1.67) |
| | 40 to 49 | 1035 | 0.39(0.35,0.44) | 0.01(0.00,0.02) | 0.16(0.13,0.19) | 0.44(0.38,0.50) | 0.98(0.78,1.18) | 1.71(1.19,2.23) |
| | 0 to <1x poverty line | 876 | 0.26(0.23,0.30) | 0.00(0.00,0.00) | 0.07(0.05,0.09) | 0.29(0.24,0.35) | 0.70(0.56,0.85) | 1.06(0.77,1.34) |
| | 1x to <2x poverty | 856 | 0.28(0.25,0.31) | 0.00(0.00,0.00) | 0.11(0.07,0.14) | 0.36(0.30,0.42) | 0.80(0.63,0.97) | 1.27(1.08,1.46) |
| Appual | 2x to <3x poverty line | 516 | 0.28(0.23,0.34) | 0.00(-0.01,0.01) | 0.11(0.08,0.13) | 0.32(0.25,0.40) | 0.77(0.55,0.99) | 1.12(0.72,1.52) |
| Annual Income | 3x to <4x poverty line | 368 | 0.33(0.25,0.41) | 0.00(-0.01,0.01) | 0.13(0.07,0.20) | 0.41(0.32,0.49) | 0.74(0.47,1.01) | 1.17(0.39,1.95) |
| Income | 4x to <5x poverty line | 251 | 0.38(0.30,0.46) | 0.01(-0.01,0.03) | 0.17(0.10,0.24) | 0.45(0.36,0.54) | 0.94(0.48,1.39) | 1.56(0.75,2.37) |
| | >= 5x poverty line | 498 | 0.46(0.39,0.52) | 0.01(-0.01,0.04) | 0.21(0.15,0.26) | 0.52(0.42,0.62) | 1.19(0.98,1.39) | 1.77(1.37,2.17) |
| | Missing/Refused/DK | 343 | 0.36(0.28,0.43) | 0.01(-0.01,0.03) | 0.14(0.11,0.17) | 0.32(0.24,0.41) | 0.79(0.40,1.17) | 1.52(0.83,2.20) |
| | <median age<="" education="" for="" td=""><td>1175</td><td>0.24(0.21,0.26)</td><td>0.00(0.00,0.00)</td><td>0.07(0.04,0.10)</td><td>0.26(0.21,0.31)</td><td>0.66(0.53,0.79)</td><td>0.99(0.83,1.16)</td></median> | 1175 | 0.24(0.21,0.26) | 0.00(0.00,0.00) | 0.07(0.04,0.10) | 0.26(0.21,0.31) | 0.66(0.53,0.79) | 0.99(0.83,1.16) |
| Education | Median education for age | 1256 | 0.33(0.29,0.37) | 0.00(0.00,0.01) | 0.14(0.12,0.17) | 0.39(0.35,0.43) | 0.79(0.64,0.93) | 1.34(1.01,1.68) |
| | >Median education for age | 904 | 0.45(0.41,0.49) | 0.02(0.00,0.04) | 0.22(0.17,0.27) | 0.53(0.46,0.59) | 1.14(0.95,1.34) | 1.90(1.47,2.33) |
| | Unknown education | 373 | 0.17(0.13,0.20) | 0.00(0.00,0.00) | 0.04(0.01,0.07) | 0.20(0.11,0.30) | 0.44(0.32,0.55) | 0.67(0.46,0.88) |

| | Times eaten in 30 days | N | Arith. Mean (95% Cl) | Selected percentiles (95% CI) | | | | | |
|-----------------|------------------------------|-----|-------------------------|-------------------------------|-------------------|------------------|-------------------|-------------------|--|
| Survey release | | | | 25th | 50th | 75th | 90th | 95th | |
| | 0 | 201 | 0.31(0.25,0.37) | 0.14 (0.13,0.14) | 0.20(0.15,0.24) | 0.30(0.20,0.40) | 0.53 (0.04,1.02) | 0.92 (0.19,1.66) | |
| | 1 | 137 | 0.61(0.40,0.82) | 0.16 (0.12,0.21) | 0.32(0.27,0.37) | 0.57(0.38,0.77) | 1.09 (0.69, 1.49) | 1.47 (.,.)ª | |
| 2013-2014 | 2 | 111 | 0.63 (0.54,0.73) | 0.27(0.19,0.35) | 0.42(0.29,0.54) | 0.72 (0.46,0.99) | 1.31(0.90,1.73) | 1.60 (0.67, 2.53) | |
| 2013-2014 | 3 | 77 | 0.62(0.54,0.70) | 0.32 (0.26,0.38) | 0.56(0.45,0.66) | 0.84 (0.66,1.02) | 1.08 (0.48,1.69) | 1.19 (.,.)ª | |
| | 4-5 | 79 | 0.86 (0.76,0.97) | 0.37 (0.26, 0.48) | 0.52(0.45,0.59) | 0.90 (0.71,1.09) | 1.66 (0.71,2.62) | 3.23 (2.21, 4.24) | |
| | 6 and up | 209 | 2.22 (1.66, 2.77) | 0.59 (0.47,0.72) | 1.24 (0.96, 1.52) | 2.64 (1.81,3.47) | 4.40 (2.84, 5.96) | 7.84 (2.69,12.99) | |
| | 0 | 197 | 0.31(0.26,0.36) | 0.15 (0.14,0.15) | 0.21(0.16,0.26) | 0.31(0.27,0.35) | 0.49(0.37,0.60) | 0.76 (0.49,1.03) | |
| | 1 | 106 | 0.42 (0.31,0.52) | 0.17 (0.13,0.21) | 0.30(0.27,0.33) | 0.46 (0.34,0.57) | 0.61(0.47,0.76) | 0.78(-0.18,1.74) | |
| 2015-2016 | 2 | 80 | 1.04 (0.55,1.53) | 0.27 (0.22,0.32) | 0.49(0.35,0.63) | 0.89(0.39,1.39) | 2.04 (-0.99,5.07) | 4.01(.,.)ª | |
| 2010-2010 | 3 | 69 | 0.71(0.52,0.90) | 0.35(0.22,0.48) | 0.51(0.30,0.71) | 0.87(0.43,1.30) | 1.09 (.,.)ª | 1.42 (.,.)ª | |
| | 4-5 | 100 | 1.00 (0.79,1.21) | 0.45 (0.37,0.54) | 0.68(0.48,0.88) | 1.11(0.69,1.52) | 2.01(0.64,3.39) | 3.01(1.36,4.67) | |
| | 6 and up | 188 | 1.77 (1.51,2.02) | 0.59 (0.39,0.78) | 1.11(0.90,1.33) | 2.42 (1.75,3.09) | 3.81(2.69,4.93) | 5.38 (3.47,7.29) | |
| | 0 | 554 | 0.34 (0.29,0.40) | 0.15 (0.14,0.15) | 0.16 (0.15,0.17) | 0.34 (0.24,0.43) | 0.61(0.42,0.80) | 1.00 (0.63,1.38) | |
| | 1 | 290 | 0.48(0.38,0.58) | 0.16 (0.15,0.17) | 0.28(0.24,0.32) | 0.53 (0.31,0.75) | 1.31(0.74,1.89) | 1.54 (1.14,1.94) | |
| 2017 March 2020 | 2 | 252 | 0.55 (0.46,0.64) | 0.21(0.16,0.25) | 0.43 (0.36,0.51) | 0.64 (0.44,0.84) | 1.11 (0.66, 1.57) | 1.34 (0.69,1.99) | |
| 2017-March 2020 | 3 | 211 | 0.78 (0.65,0.91) | 0.25 (0.18,0.32) | 0.39(0.26,0.53) | 0.95 (0.70,1.20) | 1.67 (1.31,2.03) | 2.25 (1.83, 2.67) | |
| | 4-5 | 282 | 1.06 (0.83, 1.29) | 0.33(0.25,0.40) | 0.62(0.47,0.78) | 1.16 (0.72,1.61) | 2.73 (1.90,3.56) | 3.44 (1.67,5.22) | |
| | 6 and up | 565 | 1.66 (1.47,1.85) | 0.61(0.51,0.70) | 1.08 (0.88,1.28) | 2.01(1.51,2.50) | 3.68(2.84,4.52) | 4.65 (3.24,6.06) | |

Table A.9a. Blood MeHg concentrations (µg/L), by frequency of consuming fish, by NHANES survey releases, women aged 16-49 years, NHANES 2013-March 2020

Note: Data from NHANES 2011-2012 data was not included in the current study.

^a Missing confidence interval because of stratum with single sampling unit.

| Survey | Times eaten | | Arith. Mean | Selected percentiles (95% CI) | | | | | |
|-----------|-------------|-----|-------------------|-------------------------------|-------------------|-------------------|-------------------|--------------------|--|
| release | in 30 days | N | (95% CI) | 25th | 50th | 75th | 90th | 95th | |
| 1999-2000 | 0 | 428 | 0.60 (0.50,0.71) | 0.21(0.12,0.39) | 0.36(0.27,0.47) | 0.69 (0.56,0.85) | 1.32 (1.08,1.61) | 1.74 (1.31,2.31) | |
| | 1 | 279 | 1.01(0.72,1.30) | 0.37(0.25,0.53) | 0.70(0.54,0.90) | 1.18 (0.79,1.75) | 2.12 (1.27,3.53) | 3.11 (1.69, 5.71) | |
| | 2 | 223 | 1.17 (0.93,1.40) | 0.38(0.27,0.53) | 0.78(0.58,1.04) | 1.54 (1.30,1.83) | 2.50 (1.75,3.57) | 3.55 (2.59,4.86) | |
| | 3 | 154 | 2.06 (0.26, 3.86) | 0.46 (0.16,1.33) | 0.90 (0.60,1.37) | 1.40 (0.49,4.02) | 3.77 (0.86,16.40) | 10.87 (3.09,38.27) | |
| | 4-5 | 227 | 2.27 (1.74,2.80) | 0.67(0.54,0.82) | 1.22 (0.89,1.67) | 2.83 (1.88,4.27) | 5.69(4.43,7.29) | 7.10 (5.50,9.17) | |
| | 6 and up | 326 | 3.36(2.75,3.97) | 1.02 (0.81,1.30) | 1.91(1.52,2.41) | 4.38 (3.60,5.34) | 8.33 (6.28,11.07) | 11.81(10.29,13.54) | |
| | 0 | 401 | 0.43 (0.33,0.54) | 0.16 (0.05,0.49) | 0.29 (0.21,0.41) | 0.55 (0.42,0.72) | 0.92 (0.77,1.09) | 1.19 (0.97,1.46) | |
| | 1 | 248 | 0.71(0.53,0.90) | 0.30(0.19,0.46) | 0.48(0.38,0.60) | 0.79 (0.63,1.01) | 1.52 (0.95,2.43) | 2.02(1.41,2.89) | |
| 2001 2002 | 2 | 250 | 0.84 (0.67,1.01) | 0.32(0.22,0.46) | 0.57(0.43,0.77) | 1.10 (0.85,1.42) | 1.76 (1.38, 2.25) | 2.14 (1.21, 3.76) | |
| 2001-2002 | 3 | 188 | 1.14 (0.78,1.50) | 0.37(0.23,0.59) | 0.77 (0.58,1.01) | 1.31(1.05,1.63) | 2.50 (1.33,4.70) | 3.22 (1.78,5.83) | |
| | 4-5 | 274 | 1.20 (0.99,1.42) | 0.44(0.34,0.58) | 0.85 (0.68,1.06) | 1.60 (1.24, 2.06) | 2.58 (1.96,3.39) | 3.44 (2.73,4.35) | |
| | 6 and up | 419 | 2.33 (1.92,2.75) | 0.73 (0.62,0.86) | 1.41(1.28,1.55) | 2.82 (2.53, 3.14) | 5.69 (4.27, 7.59) | 7.16 (5.04,10.18) | |
| | 0 | 365 | 0.38 (0.27,0.50) | 0.14 (0.03,0.57) | 0.25 (0.16, 0.40) | 0.44 (0.34,0.57) | 0.79 (0.57,1.10) | 1.19 (0.76,1.85) | |
| | 1 | 237 | 0.50 (0.36,0.65) | 0.15 (0.07, 0.34) | 0.33 (0.20,0.53) | 0.57 (0.38,0.85) | 1.07 (0.70, 1.63) | 1.40 (0.80,2.46) | |
| 2007 2004 | 2 | 205 | 0.65 (0.52,0.78) | 0.26(0.15,0.46) | 0.47 (0.36,0.61) | 0.77(0.64,0.93) | 1.43 (1.03,2.00) | 2.05 (1.29, 3.26) | |
| 2003-2004 | 3 | 162 | 0.89 (0.69,1.08) | 0.31(0.20,0.46) | 0.54 (0.44,0.66) | 1.13 (0.81,1.58) | 1.92 (1.37,2.68) | 3.09(2.19,4.35) | |
| | 4-5 | 241 | 1.15 (1.00,1.29) | 0.41(0.29,0.58) | 0.76(0.63,0.92) | 1.33 (1.12,1.57) | 2.57(1.81,3.65) | 4.19 (2.63,6.67) | |
| | 6 and up | 389 | 2.07(1.68,2.46) | 0.64 (0.51,0.81) | 1.21(0.99,1.48) | 2.61(1.92,3.54) | 4.36 (3.20,5.92) | 6.24 (3.62,10.77) | |
| | 0 | 433 | 0.37(0.25,0.48) | 0.11(0.03,0.39) | 0.22 (0.12,0.38) | 0.45 (0.33,0.61) | 0.77 (0.60,0.98) | 1.10 (0.83,1.46) | |
| | 1 | 248 | 0.64 (0.50, 0.78) | 0.20 (0.10,0.38) | 0.42(0.32,0.57) | 0.80 (0.63,1.01) | 1.33 (1.00,1.77) | 1.77 (1.14, 2.76) | |
| 2005 2006 | 2 | 224 | 0.82 (0.63,1.01) | 0.26 (0.16,0.41) | 0.50 (0.39,0.64) | 0.97 (0.80,1.17) | 1.61 (1.12,2.31) | 2.65 (1.25, 5.60) | |
| 2005-2006 | 3 | 173 | 1.06 (0.68, 1.45) | 0.32(0.20,0.49) | 0.62 (0.47,0.81) | 1.11 (0.64, 1.91) | 2.72 (1.28,5.79) | 3.22 (1.85, 5.60) | |
| | 4-5 | 235 | 1.28 (0.82, 1.73) | 0.40(0.26,0.61) | 0.76(0.62,0.93) | 1.45 (1.14,1.84) | 2.45 (1.16,5.16) | 3.86 (1.38,10.76) | |
| | 6 and up | 479 | 1.84 (1.61,2.08) | 0.64 (0.56,0.75) | 1.20 (1.01,1.44) | 2.38(2.00,2.83) | 4.07(3.34,4.95) | 5.77 (4.31,7.73) | |
| | 0 | 374 | 0.36 (0.25, 0.47) | 0.15 (0.05,0.41) | 0.25 (0.16,0.39) | 0.43 (0.34,0.53) | 0.67(0.54,0.83) | 0.96 (0.72,1.29) | |
| | 1 | 251 | 0.69 (0.45, 0.92) | 0.23 (0.13,0.40) | 0.43 (0.31,0.60) | 0.72 (0.49,1.06) | 1.43 (0.99,2.08) | 1.92 (1.05,3.50) | |
| 2007-2008 | 2 | 190 | 0.69(0.52,0.86) | 0.27 (0.17, 0.43) | 0.46(0.35,0.60) | 0.72(0.60,0.88) | 1.14 (0.72,1.82) | 1.89 (1.21,2.96) | |
| 2007-2008 | 3 | 136 | 0.82 (0.57,1.07) | 0.31(0.19,0.49) | 0.54 (0.42,0.69) | 0.84 (0.66,1.05) | 1.68 (0.93, 3.02) | 2.51(1.18,5.34) | |
| | 4-5 | 197 | 1.05 (0.87, 1.23) | 0.45(0.34,0.60) | 0.79 (0.62,1.01) | 1.32 (1.16,1.49) | 1.83 (1.51,2.21) | 2.34 (1.22,4.49) | |
| | 6 and up | 345 | 1.95 (1.54,2.37) | 0.64 (0.50,0.82) | 1.24 (0.94,1.62) | 2.61(2.10,3.24) | 4.21(2.96,5.99) | 6.72 (4.49,10.06) | |
| | 0 | 413 | 0.50 (0.40,0.60) | 0.19 (0.08,0.45) | 0.31(0.22,0.42) | 0.56 (0.48,0.66) | 1.01(0.73,1.39) | 1.41(1.13,1.75) | |
| | 1 | 250 | 0.58 (0.50,0.67) | 0.25 (0.16,0.40) | 0.43 (0.34,0.54) | 0.76 (0.63,0.93) | 1.19 (1.02,1.38) | 1.51(1.20,1.91) | |
| 2000 2010 | 2 | 213 | 0.81(0.67,0.94) | 0.33 (0.23,0.46) | 0.55(0.46,0.67) | 1.05 (0.76,1.46) | 1.68 (1.45,1.93) | 2.06(1.50,2.82) | |
| 2009-2010 | 3 | 132 | 0.87 (0.70,1.04) | 0.34 (0.24,0.48) | 0.56(0.43,0.73) | 1.00 (0.74,1.33) | 2.01(1.36,2.97) | 2.53 (1.96, 3.26) | |
| | 4-5 | 258 | 1.27 (1.05,1.50) | 0.47(0.37,0.60) | 0.80 (0.61,1.06) | 1.55 (1.20,2.00) | 2.60 (2.23,3.03) | 3.15 (2.29,4.35) | |
| | 6 and up | 520 | 2.11(1.87,2.35) | 0.77 (0.66,0.91) | 1.36 (1.15,1.61) | 2.72 (2.49,2.98) | 4.24 (3.54,5.08) | 6.47 (5.61,7.47) | |

Table A.9b.Blood MeHg concentrations (ug/L), by frequency of consuming fish, by NHANES survey release, women aged 16-49 years, NHANES 1999-
2010 (United States Environmental Protection Agency, 2013)

| reporting any | fish consumption in t | | | |
|------------------------|-----------------------|------------|---------|------------|
| | Parameter | Std. error | p-Value | Odds ratio |
| Intercept | 1.3154 | 0.06 | <0.0001 | |
| Income, Overall | | | 0.0004 | |
| 0 to <1x poverty line | -0.3618 | 0.09 | <0.0001 | 0.70 |
| 1x to <2x poverty line | -0.2491 | 0.10 | 0.0133 | 0.78 |
| 2x to <3x poverty line | -0.0800 | 0.11 | 0.4851 | 0.92 |
| 3x to <4x poverty line | 0.0170 | 0.19 | 0.9276 | 1.02 |
| 4x to <5x poverty line | 0.2168 | 0.17 | 0.2108 | 1.24 |
| >= 5x poverty line | 0.1546 | 0.15 | 0.3051 | 1.17 |
| Missing/Refused/DK | 0.3025 | 0.15 | 0.0430 | 1.35 |
| Race, Overall | | | <0.0001 | |
| Mexican American | 0.1636 | 0.09 | 0.0733 | 1.18 |
| Non-Hispanic Black | 0.3179 | 0.12 | 0.0081 | 1.37 |
| Non-Hispanic White | -0.5135 | 0.08 | <0.0001 | 0.60 |
| Other Hispanic | 0.1544 | 0.13 | 0.2407 | 1.17 |
| Other Race | -0.1225 | 0.12 | 0.3129 | 0.88 |
| Education, Overall | | | <0.0001 | |
| < Median for age | -0.4417 | 0.10 | <0.0001 | 0.64 |
| Median for age | 0.0740 | 0.09 | 0.4152 | 1.08 |
| >Median for age | 0.1674 | 0.11 | 0.1204 | 1.18 |
| Unknown education | 0.2003 | 0.15 | 0.1773 | 1.22 |
| Age, Overall | | | <0.0001 | |
| 16 to 19 | -0.6722 | 0.13 | <0.0001 | 0.51 |
| 20 to 29 | -0.0005 | 0.09 | 0.9953 | 1.00 |
| 30 to 39 | 0.1792 | 0.08 | 0.0182 | 1.20 |
| 40 to 49 | 0.4935 | 0.08 | <0.0001 | 1.64 |

Table A.10.Parameter estimates and odds ratio from the logistic model predicting the probability of
reporting any fish consumption in the previous 30 days, NHANES 2013-March 2020

| trequency of | fish consumption in th | | | |
|------------------------|------------------------|------------|---------|----------------|
| | Parameter | Std. error | p-Value | Relative ratio |
| Intercept | 1.3014 | 0.04 | <0.0001 | |
| Income, Overall | | | 0.1650 | |
| 0 to <1x poverty line | -0.0871 | 0.05 | 0.0754 | 0.92 |
| 1x to <2x poverty line | -0.0237 | 0.06 | 0.6931 | 0.98 |
| 2x to <3x poverty line | -0.0781 | 0.06 | 0.1681 | 0.92 |
| 3x to <4x poverty line | 0.0075 | 0.06 | 0.8975 | 1.01 |
| 4x to <5x poverty line | 0.0281 | 0.07 | 0.6920 | 1.03 |
| >= 5x poverty line | 0.1556 | 0.06 | 0.0125 | 1.17 |
| Missing/Refused/DK | -0.0023 | 0.06 | 0.9694 | 1.00 |
| Race, Overall | | | <0.0001 | |
| Mexican American | -0.0966 | 0.05 | 0.0810 | 0.91 |
| Non-Hispanic Black | 0.0889 | 0.04 | 0.0344 | 1.09 |
| Non-Hispanic White | -0.1258 | 0.04 | 0.0020 | 0.88 |
| Other Hispanic | -0.0401 | 0.06 | 0.5124 | 0.96 |
| Other Race | 0.1737 | 0.05 | 0.0008 | 1.19 |
| Education, Overall | | | <0.0001 | |
| < Median for age | -0.1865 | 0.04 | <0.0001 | 0.83 |
| Median for age | -0.0414 | 0.04 | 0.3249 | 0.96 |
| >Median for age | 0.1342 | 0.05 | 0.0146 | 1.14 |
| Unknown education | 0.0937 | 0.09 | 0.3092 | 1.10 |
| Age, Overall | | | <0.0001 | |
| 16 to 19 | -0.3314 | 0.06 | <0.0001 | 0.72 |
| 20 to 29 | 0.1203 | 0.04 | 0.0044 | 1.13 |
| 30 to 39 | 0.1582 | 0.04 | 0.0002 | 1.17 |
| 40 to 49 | 0.0529 | 0.04 | 0.1498 | 1.05 |

Table A.11.Parameter estimates and relative ratios from the model predicting the log-transformedfrequency of fish consumption in the previous 30 days (times), NHANES 2013-March 2020

| amount of fish consumed in a meal (meal size)(g), NHANES 2013-March 2020 | | | | | | | | |
|--|-----------|------------|---------|-----------------------|--|--|--|--|
| | Parameter | Std. error | p-Value | Relative ratio | | | | |
| Intercept | 4.1932 | 0.01 | <0.0001 | | | | | |
| Income, Overall | | | 0.4531 | | | | | |
| 0 to <1x poverty line | 0.0017 | 0.01 | 0.9062 | 1.00 | | | | |
| 1x to <2x poverty line | 0.0072 | 0.02 | 0.7051 | 1.01 | | | | |
| 2x to <3x poverty line | -0.0102 | 0.02 | 0.5414 | 0.99 | | | | |
| 3x to <4x poverty line | 0.0043 | 0.02 | 0.8293 | 1.00 | | | | |
| 4x to <5x poverty line | 0.0272 | 0.03 | 0.3020 | 1.03 | | | | |
| >= 5x poverty line | 0.0166 | 0.02 | 0.3624 | 1.02 | | | | |
| Missing/Refused/DK | -0.0467 | 0.02 | 0.0670 | 0.95 | | | | |
| Race, Overall | | | <0.0001 | | | | | |
| Mexican American | 0.0881 | 0.02 | <0.0001 | 1.09 | | | | |
| Non-Hispanic Black | 0.1170 | 0.01 | <0.0001 | 1.12 | | | | |
| Non-Hispanic White | 0.0923 | 0.01 | <0.0001 | 1.10 | | | | |
| Other Hispanic | -0.0247 | 0.02 | 0.1802 | 0.98 | | | | |
| Other Race | -0.2727 | 0.01 | <0.0001 | 0.76 | | | | |
| Education, Overall | | | 0.0155 | | | | | |
| < Median for age | -0.0096 | 0.01 | 0.5160 | 0.99 | | | | |
| Median for age | 0.0208 | 0.01 | 0.1598 | 1.02 | | | | |
| >Median for age | 0.0450 | 0.02 | 0.0080 | 1.05 | | | | |
| Unknown education | -0.0562 | 0.03 | 0.0958 | 0.95 | | | | |
| Age, Overall | | | 0.2898 | | | | | |
| 16 to 19 | -0.0482 | 0.03 | 0.1143 | 0.95 | | | | |
| 20 to 29 | 0.0050 | 0.01 | 0.7275 | 1.00 | | | | |
| 30 to 39 | 0.0270 | 0.01 | 0.0628 | 1.03 | | | | |
| 40 to 49 | 0.0162 | 0.01 | 0.2479 | 1.02 | | | | |

 Table A.12.
 Parameter estimates and relative ratios from the model predicting the log-transformed amount of fish consumed in a meal (meal size)(a), NHANES 2013-March 2020

| mercury conc | entration of the fish o | consumed (µg/g), NH. | ANES 2013-March 2 | 020 |
|------------------------|-------------------------|----------------------|-------------------|-----------------------|
| | Parameter | Std. error | p-Value | Relative ratio |
| Intercept | -2.9757 | 0.03 | <0.0001 | |
| Income, Overall | | | 0.4026 | |
| 0 to <1x poverty line | 0.0347 | 0.03 | 0.2938 | 1.04 |
| 1x to <2x poverty line | 0.0606 | 0.04 | 0.1782 | 1.06 |
| 2x to <3x poverty line | 0.0140 | 0.04 | 0.7537 | 1.01 |
| 3x to <4x poverty line | 0.0173 | 0.05 | 0.7131 | 1.02 |
| 4x to <5x poverty line | -0.0836 | 0.06 | 0.1897 | 0.92 |
| >= 5x poverty line | 0.0178 | 0.05 | 0.7121 | 1.02 |
| Missing/Refused/DK | -0.0609 | 0.06 | 0.3112 | 0.94 |
| Race, Overall | | | 0.2109 | |
| Mexican American | -0.0755 | 0.04 | 0.0510 | 0.93 |
| Non-Hispanic Black | 0.0201 | 0.03 | 0.5046 | 1.02 |
| Non-Hispanic White | 0.0447 | 0.03 | 0.0891 | 1.05 |
| Other Hispanic | -0.0176 | 0.03 | 0.6127 | 0.98 |
| Other Race | 0.0283 | 0.03 | 0.3708 | 1.03 |
| Education, Overall | | | 0.2515 | |
| < Median for age | 0.0432 | 0.04 | 0.2366 | 1.04 |
| Median for age | 0.0703 | 0.04 | 0.0508 | 1.07 |
| >Median for age | 0.0273 | 0.05 | 0.5543 | 1.03 |
| Unknown education | -0.1408 | 0.09 | 0.1102 | 0.87 |
| Age, Overall | | | 0.1633 | |
| 16 to 19 | -0.0711 | 0.06 | 0.2513 | 0.93 |
| 20 to 29 | -0.0322 | 0.04 | 0.3688 | 0.97 |
| 30 to 39 | 0.0428 | 0.03 | 0.1317 | 1.04 |
| 40 to 49 | 0.0605 | 0.04 | 0.0953 | 1.06 |

 Table A.13.
 Parameter estimates and relative ratios from the model predicting the log-transformed

 mercury concentration of the fish consumed (µa/a), NHANES 2013-March 2020

| Inverse of boo | dy weight (1/kg), NHAN | | | |
|------------------------|------------------------|------------|---------|-----------------------|
| | Parameter | Std. error | p-Value | Relative ratio |
| Intercept | -4.2811 | 0.01 | <0.0001 | |
| Income, Overall | | | 0.4767 | |
| 0 to <1x poverty line | -0.0143 | 0.01 | 0.2739 | 0.99 |
| 1x to <2x poverty line | -0.0016 | 0.01 | 0.9120 | 1.00 |
| 2x to <3x poverty line | -0.0182 | 0.02 | 0.3644 | 0.98 |
| 3x to <4x poverty line | -0.0126 | 0.02 | 0.5212 | 0.99 |
| 4x to <5x poverty line | -0.0031 | 0.02 | 0.8997 | 1.00 |
| >= 5x poverty line | 0.0217 | 0.01 | 0.1222 | 1.02 |
| Missing/Refused/DK | 0.0282 | 0.02 | 0.1238 | 1.03 |
| Race, Overall | | | <0.0001 | |
| Mexican American | 0.0009 | 0.01 | 0.9383 | 1.00 |
| Non-Hispanic Black | -0.1202 | 0.01 | <0.0001 | 0.89 |
| Non-Hispanic White | -0.0249 | 0.01 | 0.0151 | 0.98 |
| Other Hispanic | 0.0675 | 0.01 | <0.0001 | 1.07 |
| Other Race | 0.0767 | 0.02 | <0.0001 | 1.08 |
| Education, Overall | | | 0.0016 | |
| < Median for age | -0.0218 | 0.02 | 0.1645 | 0.98 |
| Median for age | -0.0186 | 0.01 | 0.1388 | 0.98 |
| >Median for age | 0.0465 | 0.01 | 0.0011 | 1.05 |
| Unknown education | -0.0061 | 0.03 | 0.8079 | 0.99 |
| Age, Overall | | | <0.0001 | |
| 16 to 19 | 0.0934 | 0.02 | 0.0004 | 1.10 |
| 20 to 29 | 0.0236 | 0.01 | 0.1029 | 1.02 |
| 30 to 39 | -0.0563 | 0.01 | <0.0001 | 0.95 |
| 40 to 49 | -0.0607 | 0.01 | <0.0001 | 0.94 |

Table A.14.Parameter estimates and relative ratios from the model predicting the log-transformed
inverse of body weight (1/kg), NHANES 2013-March 2020

| mercury intake per unit body weight (µg/kg), NHANES 2013-March 2020 | | | | | | | | |
|---|-----------|------------|---------|-----------------------|--|--|--|--|
| | Parameter | Std. error | p-Value | Relative ratio | | | | |
| Intercept | -1.7622 | 0.06 | <0.0001 | | | | | |
| Income, Overall | | | 0.5203 | | | | | |
| 0 to <1x poverty line | -0.0651 | 0.07 | 0.3499 | 0.94 | | | | |
| 1x to <2x poverty line | 0.0424 | 0.10 | 0.6660 | 1.04 | | | | |
| 2x to <3x poverty line | -0.0926 | 0.09 | 0.3087 | 0.91 | | | | |
| 3x to <4x poverty line | 0.0165 | 0.09 | 0.8481 | 1.02 | | | | |
| 4x to <5x poverty line | -0.0314 | 0.12 | 0.7995 | 0.97 | | | | |
| >= 5x poverty line | 0.2118 | 0.10 | 0.0435 | 1.24 | | | | |
| Missing/Refused/DK | -0.0817 | 0.10 | 0.4038 | 0.92 | | | | |
| Race, Overall | | | 0.2830 | | | | | |
| Mexican American | -0.0832 | 0.08 | 0.3079 | 0.92 | | | | |
| Non-Hispanic Black | 0.1058 | 0.05 | 0.0363 | 1.11 | | | | |
| Non-Hispanic White | -0.0137 | 0.05 | 0.7691 | 0.99 | | | | |
| Other Hispanic | -0.0150 | 0.09 | 0.8678 | 0.99 | | | | |
| Other Race | 0.0060 | 0.07 | 0.9303 | 1.01 | | | | |
| Education, Overall | | | 0.0003 | | | | | |
| < Median for age | -0.1747 | 0.06 | 0.0074 | 0.84 | | | | |
| Median for age | 0.0311 | 0.06 | 0.6178 | 1.03 | | | | |
| >Median for age | 0.2530 | 0.08 | 0.0033 | 1.29 | | | | |
| Unknown education | -0.1094 | 0.14 | 0.4509 | 0.90 | | | | |
| Age, Overall | | | 0.0035 | | | | | |
| 16 to 19 | -0.3573 | 0.10 | 0.0010 | 0.70 | | | | |
| 20 to 29 | 0.1166 | 0.06 | 0.0404 | 1.12 | | | | |
| 30 to 39 | 0.1717 | 0.06 | 0.0057 | 1.19 | | | | |
| 40 to 49 | 0.0689 | 0.06 | 0.2890 | 1.07 | | | | |

Table A.15. Parameter estimates and relative ratios from the model predicting the log-transformed mercury intake per unit body weight (µg/kg), NHANES 2013-March 2020

Appendix B

Geographic Distribution of Blood MeHg in the General U.S. Population Using NHANES 2009-2012

B1. Introduction

This analysis documents a study of the 1999-2012 National Health and Nutrition Examination Survey (NHANES) blood mercury and fish consumption data from the general U.S. population of women 16 to 49 years. Trends over time in blood methylmercury (MeHg) concentrations and demographic characteristics that are related to blood mercury concentrations have been updated and addressed with additional NHANES data 2013-March 2020 in the main report. Information presented in this appendix summarizes the analysis of the geographic distribution of blood mercury using NHANES 2009-2012 to reflect levels of recent years at the time of analysis and the blood MeHg modeling using NHANES 1999-2012.

B2. Methods

This geographic analysis applied the same methods detailed in sections 2.1 to 2.4 in the main report. In addition, serum cotinine concentration from laboratory tests and alcohol consumption data collected from questionnaires were included in this analysis. Serum cotinine served as a biomarker for assessing smoking status. The serum cotinine concentrations were log-transformed. Alcohol consumption has been shown to be related to blood mercury levels (Park and Lee, 2013). Alcohol consumption data are available for participants age 20 years and older. The analysis variable used in the model has four levels – greater than or equal to 12 drinks in past year, less than 12 drinks in past year, participants less than 20 years old, and participants with missing data for this variable.

Household annual income is derived using two sets of questions, the first for less than or greater than \$20K and the second with a more detailed breakdown. The seven income categories used for the analysis are: less than \$20K, \$20K to \$45K, greater than \$45K to \$75K, greater than or equal to \$20K (if second, more detailed question not answered), "refused" or "don't know," and missing information.

U.S. Census Region and coastal status data were obtained from the NCHS Research Data Center (http://www.cdc.gov/rdc/) through a proposal process for access to non-publicly available NHANES data. NCHS analysts derived the census region variable from the state of residence of the participant and the coastal status variable from the county of residence. If a county was adjacent to the Atlantic Ocean, Pacific Ocean, Gulf of Mexico, or one of the Great Lakes, it was considered coastal. All other counties were considered non-coastal. A list of coastal counties based on methods detailed in Mahaffey et al 2009, regardless of the geographic association of NHANES participants can be found in Table B.1. All analyses utilizing these restricted access variables were conducted at NCHS Research Data Center in Hyattsville, MD.

| State | Coastal counties | State | Coastal counties | State | Coastal counties |
|--------------------------|--|--------------------|--------------------------------------|-----------|-----------------------------------|
| Alabama | Baldwin County | Florida | Alachua County | Florida | Sarasota County |
| Alabama | Mobile County | Florida | Baker County | Florida | Seminole County |
| Alaska | Aleutians East Borough | Florida | Bay County | Florida | St. Johns County |
| Alaska | Aleutians West | Florida | Bradford County | Florida | St. Lucie County |
| Alaska | Anchorage Borough | Florida | Brevard County | Florida | Sumter County |
| Alaska | Bethel | Florida | Broward County | Florida | Suwannee County |
| Alaska | Bristol Bay Borough | Florida | Calhoun County | Florida | Taylor County |
| Alaska | City & Borough of Juneau | | 5 | | Union County |
| | | Florida | Charlotte County | Florida | |
| Alaska | City & Borough of Sitka | Florida | Citrus County | Florida | Volusia County |
| Alaska | Dillingham | Florida | Clay County | Florida | Wakulla County |
| Alaska | Haines Borough | Florida | Collier County | Florida | Walton County |
| Alaska | Kenai Peninsula Borough Ketchikan Gateway | Florida | Columbia County | Florida | Washington County |
| Alaska | Borough | Florida | DeSoto County | Georgia | Bryan County |
| Alaska | Kodiak Island Borough | Florida | Dixie County | Georgia | Camden County |
| Alaska | Lake And Peninsula Borough | Florida | Duval County/City of Jacksonville | Georgia | Chatham County |
| Alaska | Nome | Florida | Escambia County | Georgia | Glynn County |
| Alaska | North Slope Borough | Florida | Flagler County | Georgia | Liberty County |
| Alaska | Northwest Arctic | Florida | Franklin County | Georgia | McIntosh County |
| | Borough Prince of Wales-Outer | | | - | , |
| Alaska | Ketchikan | Florida | Gadsden County | Hawaii | Hawaii County |
| Alaska | Skagway-Hoonah- Angoon | Florida | Gilchrist County | Hawaii | Honolulu City and Coun |
| Alaska | Valdez-Cordova | Florida | Glades County | Hawaii | Kalawao |
| Alaska | Wade Hampton | Florida | Gulf County | Hawaii | Kauai County |
| Alaska | Wrangell-Petersburg | Florida | Hamilton County | Hawaii | Maui County |
| Alaska | Yakutat | Florida | Hardee County | Illinois | Cook |
| California | Alameda County | Florida | Hendry County | Illinois | DuPage |
| California | Contra Costa County | Florida | Hernando County | Illinois | Kane |
| California | Del Norte County | Florida | Highlands County | Illinois | Lake |
| California | Humboldt County | Florida | Hillsborough County | Illinois | McHenry |
| California | Los Angeles County | Florida | Holmes County | Illinois | Will |
| California | Marin County | Florida | Indian River County | Indiana | Lake |
| California | Mendocino County | Florida | Jackson County | Indiana | LaPorte |
| California | Monterey County | Florida | Jefferson County | Indiana | Porter |
| California | Napa County | Florida | Lafayette County | Louisiana | Assumption Parish |
| | 1 2 | Florida | , , | | |
| California | Orange County | | Lake County | Louisiana | Cameron Parish |
| California California | San Diego County San Francisco City & | Florida Florida | Lee County Leon County | Louisiana | Iberia Parish Jefferson Parish |
| Jamornia | County | TIUTIUA | Leon county | Louisiana | Lafayette Consolidated |
| California | San Luis Obispo County | Florida | Levy County | Louisiana | Government |
| California | San Mateo County | Florida | Liberty County | Louisiana | Lafourche Parish |
| California | Santa Barbara County | Florida | Madison County | Louisiana | Livingston Parish |
| California | Santa Clara County | Florida | Manatee County | Louisiana | Orleans Parish |
| California | Santa Cruz County | Florida | Marion County | Louisiana | Plaquemines Parish |
| California | Solano County | Florida | Martin County | Louisiana | St. Bernard Parish |
| California | Sonoma County | Florida | Miami-Dade County | Louisiana | St. Charles Parish |
| California | Ventura County | Florida | Monroe County | Louisiana | St. James Parish |
| Connecticut | Fairfield County | Florida | Nassau County | Louisiana | St. John The Baptist |
| Connecticut | Hartford County | Florida | Okaloosa County | Louisiana | Parish St. Mary Parish |
| Connecticut | Middlesex County | Florida | Okeechobee County | Louisiana | St. Tammany Parish |
| Connecticut | New Haven County | Florida | Orange County | Louisiana | Tangipahoa Parish |
| | , | | | | |
| Connecticut | New London County | Florida | Osceola County | Louisiana | Terrebonne Parish |
| Connecticut | Tolland County | Florida | Palm Beach County | Louisiana | Vermilion Parish |
| Connecticut | Windham County | Florida | Pasco County | Maine | Androscoggin County |
| Delaware | Kent County | Florida | Pinellas County | Maine | Cumberland County |
| Delaware | New Castle County | Florida | Polk County | Maine | Hancock County |
| Delaware | Sussex County | Florida | Putnam County | Maine | Kennebec County |
| District of Columbia | District of Columbia | Florida | Santa Rosa County | Maine | Knox County |

| State | Coastal counties | State | Coastal counties | State | Coastal counties |
|---------------|-------------------------|---------------|-------------------------|----------------|------------------------------------|
| Maine | Lincoln County | Michigan | Marquette | New York | Queens County |
| Maine | Sagadahoc County | Michigan | Mason | New York | Richmond County (Staten Island) |
| Maine | Waldo County | Michigan | Menominee | New York | Rockland County |
| Maine | Washington County | Michigan | Midland | New York | Suffolk County |
| Maine | York County | Michigan | Monroe | New York | Westchester County |
| Maryland | Anne Arundel County | Michigan | Muskegon | North Carolina | Beaufort County |
| Maryland | Baltimore City | Michigan | Oakland | North Carolina | Bertie County |
| Maryland | Baltimore County | Michigan | Oceana | North Carolina | Brunswick County |
| Maryland | Calvert County | Michigan | Ontonagon | North Carolina | Camden County |
| Maryland | Caroline County | Michigan | Ottawa | North Carolina | Carteret County |
| Maryland | Cecil County | Michigan | Presque Isle | North Carolina | Chowan County |
| Maryland | Charles County | Michigan | Saginaw | North Carolina | Craven County |
| Maryland | Dorchester County | Michigan | Sanilac | North Carolina | Currituck County |
| Maryland | Harford County | Michigan | Schoolcraft | North Carolina | Dare County |
| Maryland | Howard County | Michigan | St. Clair | North Carolina | Hyde County |
| Maryland | Kent County | Michigan | Tuscola | North Carolina | Jones County |
| Maryland | Montgomery County | Michigan | Van Buren | North Carolina | New Hanover County |
| Maryland | Prince George's County | Michigan | Washtenaw | North Carolina | Onslow County |
| Maryland | Queen Anne's County | Michigan | Washtenaw | North Carolina | Pamlico County |
| Maryland | Somerset County | Minnesota | Carlton | North Carolina | Pasquotank County |
| Maryland | St. Mary's County | Minnesota | Cook | North Carolina | Pasquotank County Pender County |
| Maryland | Talbot County | Minnesota | Lake | North Carolina | Perquimans County |
| Maryland | Wicomico County | Minnesota | St. Louis | North Carolina | Tyrrell County |
| Maryland | Worcester County | Mississippi | Hancock County | North Carolina | Washington County |
| Massachusetts | Barnstable County | Mississippi | Harrison County | Ohio | Ashtabula |
| Massachusetts | Bristol County | Mississippi | Jackson County | Ohio | Cuyahoga |
| Massachusetts | Dukes County | New Hampshire | Rockingham County | Ohio | Erie |
| | Essex County | New Hampshire | Strafford County | Ohio | Geauga |
| Massachusetts | 2 | | 2 | Ohio | - |
| Massachusetts | Middlesex County | New Jersey | Atlantic County | | Huron |
| Massachusetts | Nantucket County | New Jersey | Bergen County | Ohio | Lake |
| Massachusetts | Norfolk County | New Jersey | Burlington County | Ohio | Lorain |
| Massachusetts | Plymouth County | New Jersey | Camden County | Ohio | Lucas |
| Massachusetts | Suffolk County | New Jersey | Cape May County | Ohio | Medina |
| Michigan | Alcona | New Jersey | Cumberland County | Ohio | Ottawa |
| Michigan | Alger | New Jersey | Essex County | Ohio | Sandusky |
| Michigan | Allegan | New Jersey | Glouchester County | Ohio | Seneca |
| Michigan | Alpena | New Jersey | Hudson County | Ohio | Summit |
| Michigan | Antrim | New Jersey | Middlesex County | Ohio | Wood |
| Michigan | Arenac | New Jersey | Monmouth County | Oregon | Clatsop County |
| Michigan | Baraga | New Jersey | Ocean County | Oregon | Columbia County |
| Michigan | Bay | New Jersey | Passaic County | Oregon | Coos County |
| Michigan | Benzie | New Jersey | Salem County | Oregon | Curry County |
| Michigan | Berrien | New Jersey | Union County | Oregon | Douglas County |
| Michigan | Charlevoix | New York | Cattaraugus | Oregon | Lane County |
| Michigan | Cheboygan | New York | Cayuga | Oregon | Lincoln County |
| Michigan | Chippewa | New York | Chautauqua | Oregon | Multnomah County |
| Michigan | Delta | New York | Erie | Oregon | Tillamook County |
| Michigan | Emmet | New York | Genesee | Oregon | Washington County |
| Michigan | Genesee | New York | Jefferson | Pennsylvania | Crawford |
| Michigan | Gladwin | New York | Livingston | Pennsylvania | Erie |
| Michigan | Gogebic | New York | Monroe | Pennsylvania | Delaware County |
| Michigan | Grand Traverse | New York | Niagara | Pennsylvania | Montgomery County |
| Michigan | Houghton | New York | Onondaga | Pennsylvania | Philadelphia County |
| Michigan | Huron | New York | Ontario | Rhode Island | Bristol County |
| Michigan | losco | New York | Orleans | Rhode Island | Kent County |
| Michigan | Kalkaska | New York | Oswego | Rhode Island | Newport County |
| Michigan | Keweenaw | New York | Seneca | Rhode Island | Providence County |
| Michigan | Lapeer | New York | Wayne | Rhode Island | Washington County |
| Michigan | Leelanau | New York | Wyoming | South Carolina | Beaufort County |
| Michigan | Luce | New York | Bronx County | South Carolina | Berkeley County |
| Michigan | Mackinac | New York | Kings County (Brooklyn) | South Carolina | Charleston County |
| Michigan | Macomb | New York | Nassau County | South Carolina | Colleton County |
| | | | New York County | South Carolina | Georgetown County |

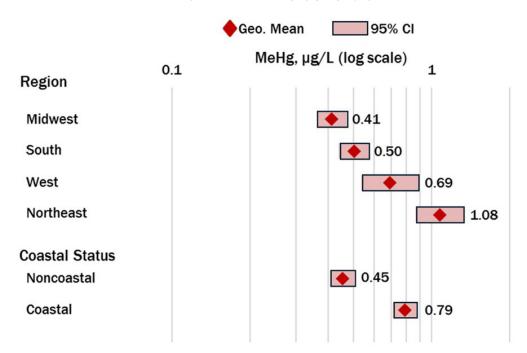
| State | Coastal counties | State | Coastal counties | State | Coastal counties |
|----------------|-------------------------|------------|-------------------------|------------|-------------------------|
| South Carolina | Horry County | Virginia | Henrico County | Washington | Jefferson County |
| South Carolina | Jasper County | Virginia | Isle of Wight County | Washington | King County |
| Texas | Aransas County | Virginia | James City County | Washington | Kitsap County |
| Texas | Brazoria County | Virginia | King and Queen County | Washington | Mason County |
| Texas | Calhoun County | Virginia | King George County | Washington | Pacific County |
| Texas | Cameron County | Virginia | Lancaster County | Washington | Pierce County |
| Texas | Chambers County | Virginia | Manassas City | Washington | San Juan County |
| Texas | Galveston County | Virginia | Manassas Park City | Washington | Skagit County |
| Texas | Harris County | Virginia | Matthews County | Washington | Snohomish County |
| Texas | Jackson County | Virginia | Middlesex County | Washington | Thurston County |
| Texas | Jefferson County | Virginia | New Kent County | Washington | Wahkiakum County |
| Texas | Kenedy County | Virginia | Newport News City | Washington | Whatcom County |
| Texas | Kleberg County | Virginia | Norfolk City | Wisconsin | Ashland |
| Texas | Matagorda County | Virginia | Northampton County | Wisconsin | Bayfield |
| Texas | Nueces County | Virginia | Northumberland County | Wisconsin | Brown |
| Texas | Orange County | Virginia | Poquoson City | Wisconsin | Calumet |
| Texas | Refugio County | Virginia | Portsmouth City | Wisconsin | Door |
| Texas | San Patricio County | Virginia | Prince William County | Wisconsin | Douglas |
| Texas | Victoria County | Virginia | Richmond City | Wisconsin | Iron |
| Texas | Willacy County | Virginia | Richmond County | Wisconsin | Kenosha |
| Virginia | Accomack County | Virginia | Stafford County | Wisconsin | Kewaunee |
| Virginia | Alexandria City | Virginia | Suffolk City | Wisconsin | Manitowoc |
| Virginia | Arlington County | Virginia | Surry County | Wisconsin | Marinette |
| Virginia | Charles City County | Virginia | Virginia Beach City | Wisconsin | Milwaukee |
| Virginia | Chesapeake City | Virginia | Westmoreland County | Wisconsin | Oconto |
| Virginia | Clifton Forge City | Virginia | Williamsburg City | Wisconsin | Ozaukee |
| Virginia | Essex County | Virginia | York County | Wisconsin | Racine |
| Virginia | Fairfax City | Washington | Clallam County | Wisconsin | Sheboygan |
| Virginia | Fairfax County | Washington | Clark County | Wisconsin | Washington |
| Virginia | Falls Church City | Washington | Cowlitz County | Wisconsin | Waukesha |
| Virginia | Gloucester County | Washington | Grays Harbor County | | |
| Virginia | Hampton City | Washington | Island County | | |

B3. Results

B3.1 Blood MeHg Summary Statistics of Geographic Distributions

Figure B-1 presents the geometric mean and 95 percent confidence interval of blood MeHg concentration by U.S. census region and coastal status for women 16 to 49 years during the survey period 2009-2012. Women in the Northeast region have the highest geometric mean concentrations while those in the Midwest have the lowest and those in coastal areas have higher geometric mean concentrations than those residing in non-coastal areas. Similar patterns are observed in the percentiles (Table B.2).

Figure B.1. Geometric mean and 95 percent Cl of MeHg by geography (NHANES 2009-2012)



| Geographie legations | N | Geometric. mean | | Percentiles | |
|----------------------|------|------------------|------|-------------|------|
| Geographic locations | N — | (95% CI) | 25th | 75th | 90th |
| Northeast | 490 | 1.08 (0.87,1.33) | 0.49 | 2.57 | 3.90 |
| Midwest | 678 | 0.41(0.36,0.47) | 0.21 | 0.72 | 1.46 |
| West | 841 | 0.69(0.54,0.90) | 0.31 | 1.51 | 2.71 |
| South | 1164 | 0.50 (0.44,0.58) | 0.24 | 1.00 | 1.94 |
| Noncoastal | 1593 | 0.45 (0.41,0.51) | 0.22 | 0.84 | 1.77 |
| Coastal | 1580 | 0.79 (0.71,0.88) | 0.35 | 1.71 | 3.25 |

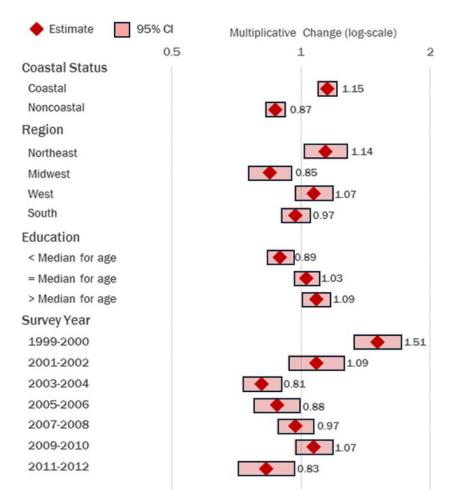
 Table B.2.
 Distribution of blood MeHg concentrations by geography (NHANES 2009-2012)

Note: Geometric mean and percentiles were calculated from the mean of 20 imputed values for each respondent.

B3.2 Blood MeHg Modeling

The results of the multivariable modeling of blood MeHg concentrations are described in this section. Transformed usual intake of mercury (TUI) through fish consumption (µg Hg/day) was the most highly significant predictor of blood MeHg concentration, with higher consumption associated with higher blood MeHg concentration. Additionally, the interaction between race/ethnicity and TUI is significant, indicating that blood MeHg concentration increases at different rates with increasing intake of fish mercury by racial/ethnic group. Other significant predictors include the geographic variables, education, and survey release. Figure B-2 presents the multiplicative change in blood MeHg concentration with 95 percent confidence intervals, for region, coastal status, and demographic characteristics. Coastal status and region are associated with blood MeHg concentration (p<0.0001 and p=0.02, respectively), with coastal residence and residence in the Northeast associated with higher blood MeHg concentrations. Residence in the Midwest is associated with lower blood MeHg concentration. Education is positively associated with blood MeHg concentrations (p=0.008). NHANES survey release is also associated with blood MeHg concentration (p<0.0001). The earliest study year, NHANES 1999-2000, has the highest concentrations and NHANES 2003-2004 and 2011-2012 have the lowest concentrations.

Figure B.2. Multiplicative change for statistically significant predictors of blood MeHg concentration (NHANES 1999-2012)



Of the remaining predictors in the model, only hematocrit, had a significant association with blood MeHg concentration (p<0.0001), with increasing hematocrit associated with increasing blood MeHg concentration. <u>Table B.3</u> provides the full model results.

| Table B.3. Modeling res | ults predictii | - | - | | | | |
|---|----------------|---------|---------|--------|---------|--------|---------|
| Parameter | Estimate | LCL | UCL | tValue | Probt | fValue | ProbF |
| Intercept | 0.6604 | 0.4314 | 0.8893 | 5.72 | <0.0001 | | |
| TUI | 1.3939 | 1.2102 | 1.5775 | 15.06 | <0.0001 | | |
| TUIV | -0.0443 | -0.1848 | 0.0962 | -0.63 | 0.5332 | | |
| TUI x log-transformed | -0.3744 | -0.7538 | 0.0051 | -1.96 | 0.0533 | | |
| body weight (centered) | -0.3744 | -0.7550 | 0.0051 | -1.50 | 0.0000 | | |
| TUI x Race/Ethnicity | | | | | | 9.06 | <0.0001 |
| TUI x Mexican American | -0.2616 | -0.4405 | -0.0828 | -2.90 | 0.0046 | | |
| TUI x non-Hispanic Black | -0.1604 | -0.3479 | 0.0271 | -1.70 | 0.0929 | | |
| TUI x non-Hispanic | 0.5524 | 0.3438 | 0.7610 | 5.25 | <0.0001 | | |
| White | 0.5524 | 0.0400 | 0.7010 | 0.20 | <0.0001 | | |
| TUI x other Hispanic | -0.0502 | -0.2981 | 0.1978 | -0.40 | 0.6891 | | |
| TUI x other race | -0.0802 | -0.3524 | 0.1920 | -0.58 | 0.5603 | | |
| Age, decade (centered) | 0.0101 | -0.0552 | 0.0754 | 0.31 | 0.7602 | | |
| Log-transformed body | -0.5116 | -1.0021 | -0.0210 | -2.07 | 0.0414 | | |
| weight (centered) | -0.5110 | -1.0021 | -0.0210 | -2.07 | 0.0414 | | |
| Log-transformed body | 0.0177 | 0 6771 | 0.070/ | 0.05 | 0.7/00 | | |
| weight2(centered) | -0.2173 | -0.6731 | 0.2384 | -0.95 | 0.3466 | | |
| Coastal Status | | | | | | 23.70 | <0.0001 |
| Coastal | 0.1395 | 0.0826 | 0.1963 | 4.87 | <0.0001 | | |
| Non-coastal | -0.1395 | -0.1963 | -0.0826 | -4.87 | <0.0001 | | |
| Region | | | | | | 3.36 | 0.0222 |
| Northeast | 0.1305 | 0.0090 | 0.2521 | 2.13 | 0.0358 | | |
| Midwest | -0.1677 | -0.2856 | -0.0498 | -2.82 | 0.0058 | | |
| West | 0.0676 | -0.0363 | 0.1716 | 1.29 | 0.2001 | | |
| South | -0.0304 | -0.1111 | 0.0502 | -0.75 | 0.4558 | | |
| Alcohol Consumption | | | | | | 2.58 | 0.0581 |
| <12 drinks past year | 0.1058 | 0.0248 | 0.1867 | 2.59 | 0.0111 | | |
| ≥12 drinks past year | 0.0264 | -0.0896 | 0.1424 | 0.45 | 0.6527 | | |
| <20 years old | -0.0958 | -0.2253 | 0.0338 | -1.47 | 0.1458 | | |
| Refused/Don't know/ | | | | | | | |
| Missing | -0.0364 | -0.2001 | 0.1273 | -0.44 | 0.6602 | | |
| Household Income | | | | | | 1.64 | 0.1458 |
| <\$20K | -0.1030 | -0.2217 | 0.0157 | -1.72 | 0.0885 | | |
| \$20K to <\$45K | -0.0348 | -0.1481 | 0.0784 | -0.61 | 0.5429 | | |
| \$45K to <\$75K | 0.0932 | -0.0214 | 0.2078 | 1.61 | 0.1102 | | |
| \$75K and over | 0.1173 | -0.0281 | 0.2626 | 1.60 | 0.1130 | | |
| \$20K and over | 0.1187 | -0.1362 | 0.3737 | 0.92 | 0.3579 | | |
| Refused/Don't know | 0.0159 | -0.3140 | 0.3458 | 0.10 | 0.9239 | | |
| Missing | -0.2072 | -0.4997 | 0.0853 | -1.41 | 0.1632 | | |
| Education | | 21.007 | 2.0000 | | 2 | 5.10 | 0.0079 |
| <median age<="" for="" td=""><td>-0.1134</td><td>-0.1847</td><td>-0.0421</td><td>-3.15</td><td>0.0022</td><td></td><td></td></median> | -0.1134 | -0.1847 | -0.0421 | -3.15 | 0.0022 | | |
| Median for age | 0.0300 | -0.0446 | 0.1046 | 0.80 | 0.4269 | | |
| >Median for age | 0.0834 | 0.0037 | 0.1632 | 2.07 | 0.0408 | | |
| Race/Ethnicity | 2,0001 | 2.0007 | | , | 2.0.00 | 4.99 | 0.0011 |
| Mexican American | -0.3862 | -0.6024 | -0.1700 | -3.54 | 0.0006 | | 0.0011 |
| Non-Hispanic Black | -0.0117 | -0.2302 | 0.2067 | -0.11 | 0.9155 | | |
| Non-Hispanic White | 0.4682 | 0.1973 | 0.7392 | 3.43 | 0.0009 | | |
| Other Hispanic | 0.0392 | -0.2748 | 0.3532 | 0.25 | 0.8048 | | |
| Other race | -0.1095 | -0.4383 | 0.2193 | -0.66 | 0.5105 | | |
| | 2.1000 | 2.1000 | 0.2100 | 0.00 | 2.0100 | | |

Table B.3. Modeling results predicting blood MeHg concentrations

| Parameter | Estimate | LCL | UCL | tValue | Probt | fValue | ProbF |
|---|----------|---------|---------|--------|---------|--------|---------|
| NHANES Survey Year | | | | | | 9.34 | <0.0001 |
| 1999-2000 | 0.4108 | 0.2809 | 0.5406 | 6.28 | <0.0001 | | |
| 2001-2002 | 0.0822 | -0.0725 | 0.2369 | 1.05 | 0.2947 | | |
| 2003-2004 | -0.2095 | -0.3164 | -0.1026 | -3.89 | 0.0002 | | |
| 2005-2006 | -0.1295 | -0.2556 | -0.0034 | -2.04 | 0.0444 | | |
| 2007-2008 | -0.0327 | -0.1316 | 0.0663 | -0.66 | 0.5140 | | |
| 2009-2010 | 0.0667 | -0.0392 | 0.1726 | 1.25 | 0.2144 | | |
| 2011-2012 | -0.1880 | -0.3420 | -0.0339 | -2.42 | 0.0174 | | |
| Log -transformed hematocrit (centered) | 1.4304 | 0.8855 | 1.9752 | 5.21 | <0.0001 | | |
| Log -transformed cotinine (centered) | 0.0298 | -0.0155 | 0.0751 | 1.31 | 0.1950 | | |
| Log -transformed cotinine2 (centered) | -0.0251 | -0.0703 | 0.0201 | -1.10 | 0.2731 | | |
| Log -transformed cotinine3 (centered) | 0.0164 | -0.0435 | 0.0763 | 0.54 | 0.5880 | | |